Contribution of Shellfish Consumption to Lower Mercury Health Risk for Residents in Northern Jiaozhou Bay, China

Lei Zhang and Lei Zhang

College of Resource and Environment, Qingdao Agricultural University, Changcheng Road 700, Qingdao 266109, China

Correspondence should be addressed to Lei Zhang; zhanglei200402@sina.com

Received 18 January 2015; Revised 9 May 2015; Accepted 12 May 2015

Academic Editor: Francesco P. Fanizzi

Copyright © 2015 L. Zhang and L. Zhang. This is an open access article distributed under the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

Fish and marine mammal consumption are an important pathway for human exposure to mercury. The low mercury content in shellfish poses a low mercury health risk to people who consume shellfish. The objectives of this study are to detect mercury concentrations in different species of shellfish and to calculate the mercury health risk from shellfish consumption among traditional residents near northern Jiaozhou Bay. A total of 356 shellfish samples, which comprised 7 species from 5 different places in northern Jiaozhou Bay, were collected from April to June in 2012. The average mercury content in the collected shellfish ranged from 0.024 mg kg$^{-1}$ to 0.452 mg kg$^{-1}$. A total of 44 shellfish samples (12.36%) had mercury levels exceeding the national pollution-free aquatic products limit (0.3 mg kg$^{-1}$). Generally, the viscus had the highest mercury content among all parts of the shellfish. A positive correlation between mercury content and total weight/edible part weight was found in most species of the collected shellfish. The results showed that shellfish consumption resulted in the lower risk of mercury exposure to residents based on the calculation of daily intake (DI) and target hazard quotient (THQ).

1. Introduction

The high level of mercury (Hg) and its compounds, which are poisonous global pollutants, are an issue of international concern [1]. Elemental, inorganic, and organic mercury, which are mainly in the form of methylmercury (MeHg), harm the environment and human health. A recent breakthrough in toxicology research on Hg indicated that the population considered safe for Hg exposure before would become the subgroup in danger of health risks from Hg exposure. Total Hg is cardiovascular, visual, and nervous system toxicity substance, and it can cause health risk to human body even under the lower Hg dose [2–4]. Moreover, the myocardial infarction, atherosclerosis, and the damage on cardiovascular system also can be caused by low-dose MeHg exposure [5, 6].

Some studies conducted in the past decades have proven that consumption of fish products and marine mammals and rice grown in contaminated paddy fields or in mining areas is the primary pathway of Hg and MeHg exposure to the general population and subpopulations living in inland areas of China, respectively [7–13]. Furthermore, some papers have investigated that the low Hg levels exist in shellfish [14–20], which suggested that the shellfish consumption may be not an important pathway to the general human population in terms of Hg exposure.

In China, shellfish aquaculture is an important component of mariculture, and in 2006, the national mariculture production reached 14.456 million tonnes, including shellfish aquaculture production of 11.136 million tonnes, which accounted for about 77% of mariculture production for that year [21]. Jiaozhou Bay is an important shellfish-breeding base in north China. The bay has more than 30 years of shellfish breeding history. Benthic aquaculture for Ruditapes philippinarum and suspended aquaculture for Argopecten irradians are the main shellfish breeding patterns in the bay. In 2006, the production of Ruditapes philippinarum in Jiaozhou Bay reached 300000 t [22]. Higher levels of Hg were then found in the intertidal zone of Jiaozhou Bay, with the Hg pollution greater in the east coast intertidal ecosystem than in intertidal zone of the west bank [23]. Traditional and frequent shellfish consumption of local residents expose them to the possible effects of Hg on their health. Therefore, investigating
the Hg content of different shellfish samples obtained from Jiaozhou Bay and assessing the Hg health risk posed by the exposure route of shellfish eating for the local population are necessary and important.

Limited literature on the subject shows that fish consumption is not an important Hg exposure pathway for coastal residents in Qingdao City, including Jiaozhou Bay [24], which is contrary to the previous conclusion from the other researches. For the traditional residents near Jiaozhou Bay, they have the habit of shellfish-eating every day, and the amount of shellfish consumption per day is more, which suggested that the shellfish-eating habit, not the fish-eating, seemed to be the Hg exposure pathway to the traditional residents in the investigated region. The conclusion in this reference “that the subpopulations living in the district of Jimo and Licang, in which the traditional residents live, had the higher hair Hg concentrations than that in the subgroups in the district of Chenyang and Laoshan, in which the foreign residents live,” seemed to support this opinion. Whether the habit of shellfish-eating causes Hg exposure risk to the residents near Jiaozhou Bay? To expand the literature, this study aims to determine the distribution of Hg content in shellfish samples among different species obtained from various locations in Jiaozhou Bay and to assess the Hg health risk resulting from the traditional shellfish consumption of coastal residents near Jiaozhou Bay. Hg content in the different parts of the shellfish and the relationship between Hg content and total weight/edible part weight of the shellfish are analyzed in this study. The results of this study provide a new understanding of Hg exposure route for residents living in the northern coastal areas in China.

2. Materials and Methods

2.1. Sample Collection. Jiaozhou Bay is located south of Shandong Peninsula in Shandong Province, northern China. From April to June in 2012, a total of 356 shellfish samples of 7 species were collected, which included Ruditapes philippinarum (n = 156, among which 101 samples collected from fishermen and 55 samples collected from remarket), Haliotis asinina (n = 21, all the samples gathered form remarket), Busycn canaliculatum (n = 38, among which 22 samples collected from fishermen and 16 samples collected from remarket), Neptunia cumingi (n = 27, among which 17 samples collected from fishermen and 10 samples collected from remarket), Concha Ostreae (n = 47, among which 35 samples collected from fishermen and 12 samples collected from remarket), Sinonovacula constricta (n = 36, among which 23 samples collected from fishermen and 13 samples collected from remarket), and Argopecten irradians (n = 31, among which 21 samples collected from fishermen and 10 samples collected from remarket), from five places in northern Jiaozhou Bay, specifically Dongdayang, Xidayang, Suliu Dock, Shaogezhuang, Houhan, and another sampling place, namely, Jiaonan, which was outside Jiaozhou Bay (Figure 1).

When we arrived at the sampling places, we collected the shellfish samples from the fishermen or at the market, with simultaneous investigation of the original location of the shellfish source. At the time of the collection, the shellfish samples were kept alive and stored in “clean” seawater to remove sediments in their digestive system, with the time of shellfish samples at the remarked before purchase being less than one day. This step aimed at retaining the delicious taste of the shellfish, which was valued by the fishermen. And then, the shellfish samples with ice placed in labelled bags were stored in foam box during their transport to the lab to keep the samples fresh. At the lab, the shellfish samples were stored in icebox at –18°C for two weeks before their detection.

2.2. Sample Preparation and Hg Determination. Shellfish samples were removed from their shells, flushed with deionized water, and blotted by using filter paper. The edible part of the shellfish was homogenized and placed into labelled bags in an icebox prior to the measurement of the Hg content. For total Hg (THg) measurement, shellfish samples (0.5–1.0 g) were digested in a sand bath by using HNO₃ and H₂SO₄ (4 : 1, v/v) catalysed by V₂O₅ and then detected by means of cold atomic absorption spectrophotometry by using an F732–V cold atomic absorption instrument and following the standard method in China [25].
The method measured by F732–V cold atomic absorption instrument has a measuring range of 0 to 10.0 μg·L⁻¹ and a sensitivity of not less than 0.05 μg·L⁻¹. All chemicals used in this experiment were produced by Sinopharm Chemical Reagent Co. Ltd., and all glassware were dipped in a mixture of HNO₃ : H₂O (1:3, v/v) for a whole night to eliminate the interference of ions attached to the glass walls. In a batch of 20 samples, three blind experiments and double parallel experiments for two shellfish samples, with their error range of less than 5%, were measured. The determination of reference materials for biological ingredients (GBW10050 (GSB–28)) was used to control the accuracy of the shellfish sample testing. Certified reference shrimp tissue (GBW–10050), with a certified value for Hg (0.049 ± 0.008 mg·kg⁻¹) issued by Reference Material Information Center of China, was used to check the performance of analytical procedure. The recovery rate in this experiment ranged from 96.3% to 103.7%.

2.3. Calculation of Daily Intake (DI) and Target Hazard Quotient (THQ). Hg intake from shellfish consumption was calculated using

\[ DI = FIR \times C, \]

where DI is Hg intake (μg·d⁻¹) and FIR is the daily intake rate for differential food (g·d⁻¹). The average national daily intake of fish and shellfish consumption is 30.1 g·d⁻¹ for adults (18 to 70 years old), 13.9 g·d⁻¹ for teenagers (13 to 17 years old), and 15.4 g·d⁻¹ for children (1 to 12 years old) in China [26]. C is the Hg concentration in the shellfish samples. In this experiment, the three shellfish species, such as Haliotis asinina, Busycon canaliculatum, and Neptunaea cumingi, used this national recommended data mentioned above in the calculation process.

A variety of shellfish is available in the market near Jiaozhou Bay. Hence, traditional residents near Jiaozhou Bay consume the cheap shellfish for meals. Our survey, which was administered near Jiaozhou Bay, asked for information on the frequency and the daily amount of shellfish consumed by residents living near Jiaozhou Bay and the price of different species of shellfish production. The results indicated that the shellfish uptake ratio was different from that mentioned above.

The average daily intakes for the four kinds of shellfish samples, such as Ruditapes philippinarum, Argopecten irradians, Sinonovacula constricta, and Concha Ostreae, were set at 50 g·d⁻¹ for adults, 23.09 g·d⁻¹ for teenagers, and 25.1 g·d⁻¹ for children on the basis of questionnaire investigation due to their large consumption per day and the cheap price. DI for people of different ages were calculated by using the mean, minimum, and maximum Hg concentrations for the different types of shellfish gathered in this experiment.

The target hazard quotient (THQ) is a method that measures human health risk from pollutant exposure [27]. THQ hypothesizes that the absorbed dose of heavy metal is equal to the intake dose and its evaluative criterion is the ratio of intake dose to reference dose. No significant health risk to the population from pollutant exposure was determined when THQ is below 1. Otherwise, the health risk exists from pollutant exposure. THQ is calculated according to

\[ THQ = \frac{(EF \times ED \times FIR \times C) \times 10^{-3}}{(RfD \times W_{AB} \times AT)}, \]

where EF is exposure frequency (365·d⁻¹), ED is exposure duration (45 years for adults, 15 years for teenagers, and 7 years for children), and FIR is the daily intake rate for differential food (g·d⁻¹, same as above). C is the Hg concentration in the shellfish samples. RfD is the reference dose (μg·kg⁻¹·bw·d⁻¹), with its PTWI value set by WHO (1972) to 0.714 μg·kg⁻¹·bw·day⁻¹. W_{AB} is the average human body weight, and it considered a weight of 60 kg for adults and teenagers and 32.7 kg for children. AT is the average time (noncarcinogenic effects AT = ED × 365·d⁻¹). ANOVA was used to compare the significant difference between the mean THQ values of different species of shellfish and of different parts of the same shellfish sample.

3. Results

3.1. Distribution of Hg Concentration in Different Species of Shellfish. Table 1 lists the Hg content of the different species of shellfish samples collected from northern Jiaozhou Bay. Among the samples, the maximum Hg value appeared in Argopecten irradians and Busycon canaliculatum at 1.072 mg·kg⁻¹ and 1.096 mg·kg⁻¹, respectively. The Ruditapes philippinarum and Busycon canaliculatum had the maximum range of Hg distribution, with their maximum Hg levels.
Table 2: The distribution of Hg in the different parts of shellfish samples (wet weight, mg kg\(^{-1}\)).

<table>
<thead>
<tr>
<th>Species</th>
<th>Name of part</th>
<th>Average</th>
<th>Range</th>
<th>S.D.</th>
<th>Number</th>
</tr>
</thead>
<tbody>
<tr>
<td>Argopecten irradians</td>
<td>Viscus</td>
<td>0.839a</td>
<td>0.523–1.072</td>
<td>0.177</td>
<td>31</td>
</tr>
<tr>
<td></td>
<td>Edible part</td>
<td>0.526b</td>
<td>0.174–0.871</td>
<td>0.273</td>
<td>31</td>
</tr>
<tr>
<td></td>
<td>Mantle</td>
<td>0.296bc</td>
<td>0.155–0.612</td>
<td>0.171</td>
<td>31</td>
</tr>
<tr>
<td></td>
<td>Foot</td>
<td>0.156bc</td>
<td>0.147–0.171</td>
<td>0.008</td>
<td>31</td>
</tr>
<tr>
<td>Concha Ostreae</td>
<td>Edible part</td>
<td>0.096a</td>
<td>0.037–0.256</td>
<td>0.067</td>
<td>47</td>
</tr>
<tr>
<td></td>
<td>mantle</td>
<td>0.047a</td>
<td>0.023–0.124</td>
<td>0.035</td>
<td>47</td>
</tr>
<tr>
<td>Busycon canaliculatum</td>
<td>Viscus</td>
<td>0.251a</td>
<td>0.011–1.096</td>
<td>0.233</td>
<td>38</td>
</tr>
<tr>
<td></td>
<td>Edible part</td>
<td>0.231a</td>
<td>0.003–0.947</td>
<td>0.219</td>
<td>38</td>
</tr>
<tr>
<td>Haliotis asinina</td>
<td>Viscus</td>
<td>0.250a</td>
<td>0.091–0.477</td>
<td>0.028</td>
<td>21</td>
</tr>
<tr>
<td></td>
<td>Edible part</td>
<td>0.121b</td>
<td>0.085–0.398</td>
<td>0.097</td>
<td>21</td>
</tr>
</tbody>
</table>

Notes: there is no significant difference between those containing same letters, and there is significant difference between those containing different letters, with \(a = 0.05\).

Table 3: Correlation between Hg content in different types of shellfish and their total weight/edible part weight.

<table>
<thead>
<tr>
<th>Species</th>
<th>Correlation coefficient (r)</th>
<th>Equation</th>
<th>Pearson coefficient ((P))</th>
<th>Number</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ruditapes philippinarum (total weight)</td>
<td>0.252*</td>
<td>(y = 0.00194x + 0.0147)</td>
<td>0.027</td>
<td>156</td>
</tr>
<tr>
<td>Ruditapes philippinarum (edible part weight)</td>
<td>0.323*</td>
<td>(y = 0.00689x + 0.01223)</td>
<td>0.004</td>
<td>156</td>
</tr>
<tr>
<td>Argopecten irradians (total weight)</td>
<td>0.452</td>
<td>(y = 0.0221x – 0.0221)</td>
<td>0.141</td>
<td>31</td>
</tr>
<tr>
<td>Argopecten irradians (edible part weight)</td>
<td>0.386</td>
<td>(y = 0.0351x + 0.0852)</td>
<td>0.215</td>
<td>31</td>
</tr>
<tr>
<td>Sinonovacula constricta (total weight)</td>
<td>0.480</td>
<td>(y = 0.0072x + 0.0011)</td>
<td>0.161</td>
<td>36</td>
</tr>
<tr>
<td>Sinonovacula constricta (edible part weight)</td>
<td>0.294</td>
<td>(y = 0.0063x + 0.0428)</td>
<td>0.409</td>
<td>36</td>
</tr>
<tr>
<td>Haliotis asinina (total weight)</td>
<td>0.938**</td>
<td>(y = 0.01672x – 0.25966)</td>
<td>0.000</td>
<td>21</td>
</tr>
<tr>
<td>Haliotis asinina (edible part weight)</td>
<td>0.948**</td>
<td>(y = 0.02183x – 0.16669)</td>
<td>0.000</td>
<td>21</td>
</tr>
<tr>
<td>Busycon canaliculatum (total weight)</td>
<td>0.908**</td>
<td>(y = 0.0072x – 0.05267)</td>
<td>0.000</td>
<td>38</td>
</tr>
<tr>
<td>Busycon canaliculatum (edible part weight)</td>
<td>0.925**</td>
<td>(y = 0.01552x – 0.01723)</td>
<td>0.000</td>
<td>38</td>
</tr>
<tr>
<td>Concha Ostreae (total weight)</td>
<td>0.381</td>
<td>(y = 0.0019x + 0.0216)</td>
<td>0.073</td>
<td>47</td>
</tr>
<tr>
<td>Concha Ostreae (edible part weight)</td>
<td>0.378</td>
<td>(y = 0.0082x + 0.0317)</td>
<td>0.076</td>
<td>47</td>
</tr>
<tr>
<td>Neptunea cumingi (total weight)</td>
<td>0.671*</td>
<td>(y = 0.00267x – 0.03605)</td>
<td>0.024</td>
<td>27</td>
</tr>
<tr>
<td>Neptunea cumingi (edible part weight)</td>
<td>0.693*</td>
<td>(y = 0.0092x – 0.00839)</td>
<td>0.018</td>
<td>27</td>
</tr>
</tbody>
</table>

** Remarkably significant correlation; * significant correlation.

being 940-fold and 365-fold higher than the minimum, respectively. A significant difference existed in the mean Hg value of the different shellfish species, which was remarked in Table 1. The results of all the kinds of shellfish had 2 groups, one with mean concentrations not exceeding 0.24 mg kg\(^{-1}\), and other with values greater than 0.24 mg kg\(^{-1}\), with Argopecten irradians having the highest mean (0.452 mg kg\(^{-1}\)).

3.2. Comparison of Hg Content in Different Parts of Shellfish. Argopecten irradians shows significant differences in Hg contents between the different parts studied (Table 2). The mean Hg value of the viscus in Argopecten irradians was remarkably higher than that of the edible part, foot, and mantle (\(p < 0.01\)). The average Hg level in the edible part was significantly higher than that in the foot and mantle (\(p < 0.01\)). For Haliotis asinina, the average Hg value in viscus was observably higher than that in edible part (\(p < 0.05\)).

3.3. Relationship between Hg Content in Shellfish and Their Weight. Table 3 indicates the correlation between Hg contents in shellfish samples and their total weigh and edible part weight. Ruditapes philippinarum and Neptunea cumingi showed significant positive relationships between Hg contents in shellfish and their total weight and edible part weight (\(p < 0.05\)). Haliotis asinina and Busycon canaliculatum indicated remarkably significant positive relationship between Hg concentrations in shellfish and their total weight and edible part weight (\(p < 0.01\)). Inversely, the residual species of shellfish samples did not show any correlation between the Hg content in shellfish samples and their total weight or edible part weight.

3.4. Intakes and THQ of Hg through Shellfish Consumption. Table 4 presents the DI of Hg based on the consumption of different species of shellfish by the traditional residents according to different age groups. The average DI for all species of shellfish was less than the daily Hg intake given by FAO/WHO, with its value being 40 \(\mu g\) d\(^{-1}\) [28]. The maximum and minimum DI appeared in the consumption of Argopecten irradians and Ruditapes philippinarum by the residents in all age groups, respectively. Only the maximum DI through Argopecten irradians consumption...
After Hg entered into an aquatic ecosystem, Hg transformed into MeHg affecting by the pollution-free aquatic products (0.3 mg kg\(^{-1}\)) study exceeded the standard limit for the Hg content in shellfish samples in this study. However, the average Hg content in shellfish in this experiment was only 1.48%–40.72% to that in other fishery products, such as tuna and swordfish collected in Japan (1.11–1.82 mg kg\(^{-1}\)) and Spain (0.470–0.540 mg kg\(^{-1}\)) [20, 26]. The over standard rate was obtained by Argopecten irradians, Haliotis asinina, and Busycon canaliculatum, which was collected from four sampling places (Dongdayang, Jiaonan, Shaogezhuang, and Houhan), was the most widely distributed shellfish species among all samples. The analysis results of the significant difference in the average Hg content of Ruditapes philippinarum from the above four places show that the Hg content in Ruditapes philippinarum from Dongdayang was significantly higher than those from Houhan (\(p < 0.05\)) and Jiaonan. Jiaozhou Bay is a semiclosed gulf, with the average and maximum water depth being 7 m and 64 m, respectively. The slow exchange velocity of seawater in this bay may lead to the easy accumulation of pollutants in the sediment. The heavy metals in the sediments of Jiaozhou Bay come from terrestrial and marine dual sources. The increase in heavy metal concentration in the seawater of Jiaozhou Bay continuously results from the increase in terrigenous pollution sources and the rapid development of the local economy. In aquatic ecosystem, the sediments usually act as a sink and source for Hg, in which Hg rereleases into the surrounding water inflecting by the biological, physical, and chemical factors [32]. After Hg entered into an aquatic ecosystem, Hg transformed into MeHg affecting by the

4. Discussion

The Hg content in 12.36% of the shellfish samples in this study exceeded the standard limit for the Hg content in pollution-free aquatic products (0.3 mg kg\(^{-1}\)) in China [29]. Moreover, the maximum and average Hg content of the shellfish samples were 2.76- to 91.33-fold and 4.96 to 150.67 times higher than those of the shellfish collected in the coastal areas of the Fujian Province (0.002–0.064 mg kg\(^{-1}\)) and in Brazil (0.206–0.397 mg kg\(^{-1}\)), Italy (0.023–0.100 mg kg\(^{-1}\)), and Spain (0.003–0.019 mg kg\(^{-1}\)), respectively [15–20, 30, 31]. However, the average Hg content in shellfish in this experiment was only 1.48%–40.72% to that in other fishery products, such as tuna and swordfish collected in Japan (1.11–1.82 mg kg\(^{-1}\)) and Spain (0.470–0.540 mg kg\(^{-1}\)) [20, 26]. The over standard rate was obtained by Argopecten irradians, Haliotis asinina, and Busycon canaliculatum, which was collected from four sampling places (Dongdayang, Jiaonan, Shaogezhuang, and Houhan), was the most widely distributed shellfish species among all samples. The analysis results of the significant difference in the average Hg content of Ruditapes philippinarum from the above four places show that the Hg content in Ruditapes philippinarum from Dongdayang was significantly higher than those from Houhan (\(p < 0.05\)) and Jiaonan. Jiaozhou Bay is a semiclosed gulf, with the average and maximum water depth being 7 m and 64 m, respectively. The slow exchange velocity of seawater in this bay may lead to the easy accumulation of pollutants in the sediment. The heavy metals in the sediments of Jiaozhou Bay come from terrestrial and marine dual sources. The increase in heavy metal concentration in the seawater of Jiaozhou Bay continuously results from the increase in terrigenous pollution sources and the rapid development of the local economy. In aquatic ecosystem, the sediments usually act as a sink and source for Hg, in which Hg rereleases into the surrounding water inflecting by the biological, physical, and chemical factors [32]. After Hg entered into an aquatic ecosystem, Hg transformed into MeHg affecting by the

### Table 4: Hg intake to residents of different age groups near northern Jiaozhou Bay by shellfish consumption values (\(\mu g\ d^{-1}\)).

<table>
<thead>
<tr>
<th>Species</th>
<th>Average</th>
<th>Range</th>
<th>Average</th>
<th>Range</th>
<th>Average</th>
<th>Range</th>
</tr>
</thead>
<tbody>
<tr>
<td>Adults</td>
<td>1.200</td>
<td>0.050–47</td>
<td>0.554</td>
<td>0.023–21.705</td>
<td>0.602</td>
<td>0.025–23.594</td>
</tr>
<tr>
<td>Children</td>
<td>4.000</td>
<td>1.650–8.550</td>
<td>1.847</td>
<td>0.762–3.948</td>
<td>2.008</td>
<td>0.828–4.292</td>
</tr>
</tbody>
</table>

### Table 5: THQ of Hg in the different species of shellfish from northern Jiaozhou Bay.

<table>
<thead>
<tr>
<th>Species</th>
<th>THQ through shellfish consumption for residents of different age</th>
<th>Adults</th>
<th>Range</th>
<th>Average</th>
<th>Range</th>
<th>Average</th>
<th>Range</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ruditapes philippinarum</td>
<td>0.028</td>
<td>0.001–1.097</td>
<td>0.013</td>
<td>0.001–0.507</td>
<td>0.026</td>
<td>0.001–1.010</td>
<td></td>
</tr>
<tr>
<td>Argopecten irradians</td>
<td>0.527</td>
<td>0.172–1.251</td>
<td>0.244</td>
<td>0.079–0.572</td>
<td>0.486</td>
<td>0.158–1.152</td>
<td></td>
</tr>
<tr>
<td>Sinonovacula constricta</td>
<td>0.093</td>
<td>0.039–0.199</td>
<td>0.043</td>
<td>0.018–0.092</td>
<td>0.086</td>
<td>0.036–0.184</td>
<td></td>
</tr>
<tr>
<td>Haliotis asinina</td>
<td>0.170</td>
<td>0.060–0.334</td>
<td>0.079</td>
<td>0.028–0.155</td>
<td>0.160</td>
<td>0.056–0.315</td>
<td></td>
</tr>
<tr>
<td>Busycon canaliculatum</td>
<td>0.176</td>
<td>0.002–0.770</td>
<td>0.081</td>
<td>0.000–0.356</td>
<td>0.165</td>
<td>0.002–0.723</td>
<td></td>
</tr>
<tr>
<td>Concha Ostreae</td>
<td>0.111</td>
<td>0.027–0.299</td>
<td>0.051</td>
<td>0.012–0.138</td>
<td>0.102</td>
<td>0.025–0.275</td>
<td></td>
</tr>
<tr>
<td>Neptuinea cumingi</td>
<td>0.051</td>
<td>0.005–0.010</td>
<td>0.024</td>
<td>0.002–0.065</td>
<td>0.048</td>
<td>0.004–0.133</td>
<td></td>
</tr>
</tbody>
</table>

*Table 4: Hg intake to residents of different age groups near northern Jiaozhou Bay by shellfish consumption values (\(\mu g\ d^{-1}\)).

*Table 5: THQ of Hg in the different species of shellfish from northern Jiaozhou Bay.*
various types of anaerobic microorganisms in the sediments, such as sulfate reducing bacteria, iron reducing bacteria, and methanogenic bacteria [33, 34]. The ratio of MeHg/THg in sediments usually ranged 0.1–2.4%, with the special samples reaching to 10% in sediments from some lakes and wetland [35–38]. After heavy metals contaminate the sediments, the chemicals that are released back into the overly seawater under suitable conditions can cause secondary pollution [39].

The growing shellfish in the sediments absorbs heavy metals by filter-feeding food, seawater, and direct contact with the surrounding sediments. This action may be the reason why the Hg content in the shellfish from the northern Jiaozhou Bay exceeded the Hg level in other places.

In this experiment, the mean Hg content in Argopecten irradians was remarkably higher than that in the residual shellfish. Compared with other bivalves, such as Ruditapes philippinarum, Sinonovacula constricta, and Neptunia cumingi, the larger size of Argopecten irradians resulted in higher Hg accumulation in their bodies, which also agreed by fish, namely, the size or age of fish, and their trophic position significantly affected the biomagnification of Hg in fish [40–42]. For Argopecten irradians, the organic detritus and plankton are their important food. Compared with water, the eating habit is the important pathway of Hg exposure for the organism in the aquatic ecosystem [43]. The chemical speciation of Hg may be an important factor inflecting Hg uptake for the aquatic organisms. In the part of Hg uptake in the organisms, HgCl₂ and CH₃HgCl were considered as the main species for Hg accumulation of Hg(II) and MeHg, respectively [44]. Moreover, the particulate matter is a transfer mechanism of Hg in water column to the sediment in an aquatic ecosystem [45]. Hg, major in MeHg, can biomagnify along the food chain, including the plankton, in an aquatic ecosystem [46, 47]. The planktonic organisms provide the essential nutrients to the higher trophic organisms, and at the same time, the pollutant, such as Hg, also is conveyed to them. And then, MeHg content in plankton increased with the increase of their size [48]. The dissolved organic carbon (DOC) was an important factor affecting the higher MeHg/THg ratio in water than that in periphyton, flocculent material and soil in everglades [49]. The inorganic Hg and monomethylmercury were strongly tied to organic matters, which is the important food for the shellfish, such as Argopecten irradians. When the organic matter content in sediment and water were rich, the amount of Hg accumulation in benthic invertebrates was lower and it seemed that monomethylmercury had the higher uptake rate to the benthic organism than inorganic Hg [50]. For shellfish samples, their Hg content in viscus was significantly higher than that in edible part (Table 2), which resulted from the filter action by gall and storage in viscus. For the benthic organisms living in the sediment, the factors for Hg accumulation in their bodies are complex.

Haliothis asininae has a longer life period of about 1 to 2 years, which could result in the Hg distribution in their edible part and shell equally. This quality could be explained partly by the positive correlation between Hg values and their edible part weight and total weight (Table 3). Generally, the proportion of the edible part to total weight increases gradually along with the age growth during the culture cycle of 1 to 3 years, with the growth speed of the edible part faster than that of the shell. For the Bivalves, such as Argopecten irradians and Concha Ostreae, they have a shorter aquaculture period of about six to eight months, and the larger weight of the shell compared with its edible part may be the reason for the noncorrelation between Hg content in the edible part and its total weight/edible part weight. For shellfish, such as Argopecten irradians, food residue in the gestive gland belonging to the viscus may be the main reason for the higher Hg content in this part than that in other parts.

In this study, two methods, namely, DI, and THQ, were used to estimate the Hg intake through shellfish consumption of the traditional residents of different age groups near Jiaozhou Bay. The calculation of DI was only based on the data of the daily Hg intake and Hg content in shellfish. Compared with Hg intake from fish and shellfish for Japanese people (24.28 µg per capita per day) [26], Hg intake from shellfish samples in this experiment was 2.28–93.08% to that data mentioned above. Only for the average Hg intake from the consumption of Argopecten irradians, it was close to Hg intake from fish and shellfish from Japan. If we consider a weight of 60 kg for adults and a PTWI for total Hg of 5 µg kg⁻¹ bw·week⁻¹ (0.714 µg·kg⁻¹ bw·day⁻¹), we have a DI of 42.84 µg·day⁻¹ for adults and teenagers. For children, considering a weight of 32.7 kg, it will be 23.35 µg·day⁻¹. The results of calculation of THQ showed that the average Hg intake per capita per day was 1.29–52.75% and 2.58–48.59% for the subgroup of adults/teenagers and children, respectively. The fact indicated that there was a lower Hg health risk from shellfish product for all the residents, including the sensitive crowd children, living near Jiaozhou Bay. However, Argopecten irradians captured in this experiment presenting the highest Hg level in all the shellfish samples also had the highest daily Hg intake, which was 1.15-fold higher than the corresponding data for children. In the long term, it seems to be safe for the residents (including children) from Hg exposure through shellfish consumption basic on the kinds and producing areas in this experiment, which also agreed by the result of the calculation of THQ. The outcome of THQ suggested that the result of THQ was 1.3–52.7% and 2.6–48.6% to the standard value of 1 for the subgroup of adults/teenagers and children, respectively. For all the subgroup, the results of the two calculation methods suggested that the long period of consumption for all kinds of shellfish in this experiment is safe for health risk from Hg intake. However, the long term consumption of Ruditapes philippinarum and Argopecten irradians seems to have a slightly higher Hg health risk to the subgroups of adults and children due to their maximum Hg concentrations. Therefore, the children, as the sensitive subpopulation, should reduce their shellfish-eating amount for the safe of Hg intake during a long period.

5. Conclusions

The Hg concentration of 12.43% found in the shellfish samples exceeded the national limit for pollution-free aquatic products. The Hg distribution in shellfish had a difference among these species of shellfish. The traditional residents
near Jiaozhou Bay had low Hg intake through shellfish consumption based on the DI and THQ calculation. And then, these residents, for all the different age subpopulations, had lower Hg intake through shellfish consumption according to the estimation from the two methods.

**Conflict of Interests**

The authors declare that there is no conflict of interests regarding the publication of this paper.

**Acknowledgment**

This research was financed by the National Natural Science Foundation of China (41101094, 4110472).

**References**


