Research Article

The Species Richness of Vascular Plants and Amphibia in Major Plant Communities in Temperate to Tropical Australia: Relationship with Annual Biomass Production

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Aerodynamic fluxes (frictional, thermal, evaporative) in the atmosphere as it flows over and through a plant community determine the Foliage Projective Covers and eco-morphological attributes of new leaves developed annually in overstorey and understorey strata. The number of leaves produced on vertical foliage shoots depends on available soil water and nutrients, also ambient temperature, during this short growth season. Stem density (number of stems per hectare) and species richness (number of species per hectare) in the overstorey of major Floristic Groups are correlated with annual shoot growth (ASG, t ha\(^{-1}\)) in that stratum. Species richness in the overstorey increases in the climatic gradient from the arid to the humid zone as well as with increasing air temperatures (about 10\(^{\circ}\)C) from temperate to tropical Australia. Species richness in the understorey is highest in plant communities in temperate Australia, decreasing in the temperature gradient towards the tropics. As with other major plant and animal groups within an ecosystem, the species richness of Amphibia is correlated with the amount of solar energy fixed (per annum) by the major plant formation in the region—a photosynthetic potential determined by the foliage shoots (ASG, t ha\(^{-1}\)) produced annually in the overstorey.

1. Introduction

Over the last 60 years, the first author has pioneered the study of community-physiology (known as crop-physiology in agriculture)—the physicochemical processes that determine the structure, growth, and biodiversity in evergreen plant communities (with associated consumers and decomposers) from the tropical north to the temperate south of Australia.

The discipline of community-physiology was promoted by the International Biological Program (IBP) as part of Section PP Physiological Processes in Terrestrial Ecosystems [1].

The structures of plant formations—closed-forest, tall open-forest, open-forest, woodland, low woodland, tall shrubland, low shrubland—that are found from the perhumid to the arid climatic zones from the tropical north to the temperate south of the Australian continent are essentially similar [2].

The Foliage Projective Cover (FPC) of the overstorey (FPCo > 2 m) and understorey (FPCu < 2 m) strata of terrestrial evergreen plant communities had to be precisely measured using upward and downward crosswire sighting tubes at every 50 cm along six randomly located line transects [3]. The combined sum of (FPCo + FPCu) has been found to be essentially constant throughout the life cycle of a “climax” plant community both in post-fire successions and in plant communities along both edaphic and wetland continua [4–7].
The Foliage Projective Covers of the overstorey and understorey strata—the horizontal spread of each plant that forms the plant formation—are determined during the short growing season of each plant within the plant community. Aerodynamic fluxes (frictional, thermal, evaporative) “abrade” the edges of each plant as the atmosphere flows over and through the plant community.

During the same short period of foliage growth, aerodynamic fluxes determine the ecomorphological development of all new leaves on vertical shoots—Leaf Area (LA, mean area of a mature leaf) and Leaf Specific Weight (LSW, dry weight per unit leaf area, a measure of the proportion of structural to cytoplasmic components of the leaf).

The number of leaves produced on each foliage shoot, however, depends on the transport of available soil nutrients and soil water in the transpiration stream during this short period of foliage growth [8, 9]. The total leaf area (per hectare) of each stratum is termed the Leaf Area Index (LAI) of the plant community.

During this short period of vertical foliage growth, essentially the same number of mature leaves from the previous year is abscissed as they are formed in the current year (demonstrated on tagged shoots by [10–15]). Hence, the previous year is abscissed as they are formed in the current year essentially the same number of mature leaves from the plant community.

The total leaf area (per hectare) of each stratum is termed the Leaf Area Index (LAI) of the plant community. The total leaf area (per hectare) of overstorey and understorey strata in plant communities (see [16, Figure 17.14a]). The differences in slope of the linear regressions in tropical and temperate Australia were integrated into one linear regression when the species richness of small mammals was plotted against the species richness of the plants recorded in only the understorey stratum [16, Figure 17.14b].

A similar statistical relationship was observed between the number of species of resident birds and the number of plant species—mostly overstorey species—recorded in remnant stands of subtropical rainforests in north-eastern New South Wales [16, 24, Figure 17.1].

The species richness of vertebrates (amphibia, lizards, mammals, birds, but not elapid snakes), also plant species, recorded in plant communities—from the arid to the humid climatic zones— in south-western Australia is consistently lower than recorded in plant communities in south-eastern Australia (see [16, 17]—summarized in Figure 1). The infertility of the ancient soils of Western Australia [20, 25], when compared with the relatively recent soils in eastern Australia, appears to have contributed to reduced species richness (per plant community) within south-western ecosystems (Figure 2).

In this paper, the relationship of species richness of overstorey and understorey plants in major plant communities, and Amphibia in associated wetlands, is explored (1) along the climatic gradient from the arid to the humid zones in temperate, subtropical, and tropical Australia and (2) in the edaphic gradient from extremely nutrient-poor to nutrient-rich soils.

2. Methods

2.1. Evaporative Coefficients of Each Sampling Locality. The community-physiological constant—the evaporative coefficient \( k \)—of each sampling locality was computed by optimising the monthly rate of usage of soil water so that some water (albeit minimal) is available for evapotranspiration from the evergreen plant community even during the driest season of the year (See (1), after Specht [26]). The perhumid climatic zone is defined, where \( k > 0.75 \times 10^{-2} \); humid climatic zone—\( k = 0.75 - 0.55 \times 10^{-2} \); subhumid climatic zone—\( k = 0.55 - 0.45 \times 10^{-2} \); semiarid climatic zone—\( k = 0.45 - 0.35 \times 10^{-2} \); arid climatic zone—\( k < 0.35 \times 10^{-2} \) (All \( k \) values per mm of soil water available per month).

For every macroclimate, in spite of seasonal and yearly fluctuations in rainfall, the monthly moisture index (M.I.), the ratio of actual to potential evapotranspiration \( \left( \frac{E_o}{E_p} \right) \), is correlated to available water \( W \) mm) by the evaporative coefficient \( k \) [26–28]—defined by analysis of long-term soil moisture studies of representative plant communities from the arid to the humid climatic zone [29–34] as

\[
\text{M.I.} = \left( \frac{E_o}{E_p} \right) \cdot \left( \frac{E_o}{E_a} \right) = kW = k(P - R - D - S_{ext}),
\]

where \( k \) is evaporative coefficient (a community-physiology constant), \( E_a \) is actual evapotranspiration (mm), \( E_p \) is potential evapotranspiration (mm), \( E_o \) is pan evaporation (mm), which was estimated using mean maximum temperature and
Figure 1: (a) Alpha diversity of amphibians, elapid snakes, lizards, and small mammals resident in plant communities in south-western (○) and south-eastern Australia (△), plotted along the climatic gradient from the semiarid (Evaporative Coefficient, $k = 0.35 - 0.45 \times 10^{-2}$) to the humid ($k = 0.55 - 0.75 \times 10^{-2}$) climatic zones (after [15, 16]). (b) Alpha diversity of birds resident in plant communities in south-western (○) and south-eastern Australia (△), plotted along the climatic gradient from the semiarid (Evaporative Coefficient, $k = 0.35 - 0.45 \times 10^{-2}$) to the humid ($k = 0.55 - 0.75 \times 10^{-2}$) climatic zones (after [15, 16]). (Data for south-western Australia are summarized by Specht in [17] and revised by Professor Martin Cody, UCLA).
The Evaporative Coefficient \((k)\) of each sampling locality thus enables an estimate of the annual growth of foliage shoots in the overstorey stratum to be computed—for plant communities on medium-nutrient soils.

### 2.3. Primary Productivity of Plant Communities along Soil Nutritional Gradients

An assessment of the influence of soil fertility on biomass production of plant communities was made (1) in nutritional experiments conducted in the field [36, 38] and (2) in glasshouse experiments conducted on the relative growth of eucalypts dominant on soils with a fertility range from extremely nutrient poor to nutrient rich—differentiated by foliar nutrient (particularly phosphorus and nitrogen) levels in overstorey leaves [16].

Relative to the annual foliage growth (\(\text{ASG, t ha}^{-1}\)) in “climax” plant communities on medium-nutrient soils (Figure 3(a)), annual shoot growth (dry weight measured in \(\text{t ha}^{-1}\)) in major plant formations on nutrient-poor soils is about 60\% less, 50\% less on extremely nutrient-poor soils, and 130\% more on nutrient-rich soils (Figure 3(a), after [16, 36]).

Details of the nutrient status of the soils associated with each Floristic Group at each collecting locality were collated from soil and vegetation surveys of Australia [16, 20, 25, 39]. These data on soil nutrient status enabled an assessment of the relative primary production of foliage shoots (\(\text{ASG, t ha}^{-1}\)) at each sampling site (Table 1).

### 2.4. Species Richness of Vascular Plants (per Hectare)

The species richness of vascular plants in any plant community may be assessed by the asymptote of the species-area curve [94], where the number of species is recorded in random quadrats of increasing area from zero to at least one hectare—essentially the area of vegetation assessed in the relevé phytoécologue in the survey of the vegetation around the Mediterranean Basin [95, 96] and in the survey of the land systems in northern Australia [97, 98].

The floristic composition of plant communities has been recorded by plant ecologists in relevés in every 30’ latitude \(\times 30’\) longitude grid cell throughout Australia. These lists have been collated in eight large data banks and analysed objectively by the classificatory program TWINSPAN [99] into Floristic Groups [39].

The numbers of vascular plants that have been recorded in the overstorey and understorey strata in the “climax” Floristic Group representative of each amphibian sampling area in southern and northern Australia are tabulated in Table 1.

As the number of months when the rooting zone is waterlogged increases along a wetland continuum, the species richness of the plant communities shows a continuous decline from the “climax” plant community [7]. The sum of overstorey and understorey Foliage Projective Covers (\(\text{FPCo} + \text{FPCu}\)), however, remains relatively constant along the wetland continuum [7].

### 2.5. Species Richness of Amphibia (per Major Plant Community)

In the 1980s, the International Society for Mediterranean Ecosystems endeavoured to collate the species

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**Figure 2:** Number of TWINSPAN Floristic Groups (per 1° latitude \(\times 1°\) longitude grid cell), plotted against the number of Soil Groups recorded in the same grid cells in south-eastern and Western Australia [19, 20]. South-eastern Australia Floristic Groups = 1.34 Soil Groups + 3.83 \((n = 15; r^2 = 0.96)\). Western Australia Floristic Groups = 0.92 Soil Groups + 1.51 \((n = 15; r^2 = 0.95)\).
Table 1: Number of species of Amphibia in plant formations in south-eastern Australia, South Australia, south-western Australia, and Northern Territory. The Foliage Projective Covers of the combined overstorey and understorey strata in each plant community determine the Evaporative Coefficient ($k$)—the ratio of actual to potential evapotranspiration per mm water available per month, a constant throughout the year, throughout the life cycle of the plant community, and between microhabitats in the same climatic area [5, 16, 26, 27, 40]. The annual shoot growth (dry weight per unit area of land) of each plant community has been estimated for each plant community [11], adjusted for soil nutrient status [16, 36].

<table>
<thead>
<tr>
<th>Plant formation</th>
<th>Locality</th>
<th>Mean annual temperature (°C)</th>
<th>Annual precipitation (mm) (runoff%)</th>
<th>Evaporative coefficient ($k \times 10^{-2}$ per mm H$_2$O)</th>
<th>Overstorey &amp; understorey spp. per hectare</th>
<th>Annual shoot growth (t ha$^{-1}$)</th>
<th>Vegetation survey</th>
<th>Amphibian survey</th>
</tr>
</thead>
<tbody>
<tr>
<td>South-eastern Australia (medium-nutrient soils∗ — clayey podsols, etc.)</td>
<td></td>
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<tr>
<td>(Tall) Open-forest</td>
<td>Nowa Nowa S.F., Vic. (37° 41'S, 148° 06'E)</td>
<td>13.7</td>
<td>865 (9.5%)</td>
<td>0.75</td>
<td>6/46</td>
<td>5.5</td>
<td>Gullan et al. [41, 42]</td>
<td>17</td>
</tr>
<tr>
<td></td>
<td>Myall Lakes, N.S.W. (32° 26'S, 152° 24'E)</td>
<td>17.5</td>
<td>1310 (72%)</td>
<td>0.70</td>
<td>5/46</td>
<td>5.0</td>
<td>Osborn and Robertson [44]</td>
<td>16</td>
</tr>
<tr>
<td></td>
<td>Nadgee N.R., N.S.W. (37° 26'S, 149° 55'E)</td>
<td>15.0</td>
<td>920 (12%)</td>
<td>0.60</td>
<td>4/44</td>
<td>3.5</td>
<td>Adams and Craven [46], Austin and Sheaffe [47], Audas [49], Frankenberg [50], Frood and Calder [51]</td>
<td>12</td>
</tr>
<tr>
<td></td>
<td>The Grampians, Vic. (37° 08'S, 142° 26'E)</td>
<td>&lt;15.0</td>
<td>914 (12%)</td>
<td>0.59</td>
<td>4/44</td>
<td>3.8</td>
<td>tyler [52]</td>
<td></td>
</tr>
<tr>
<td>Open-scrub</td>
<td>Wyperfield N.P., Vic. (35° 35'S, 142° 00'E)</td>
<td>16.5</td>
<td>344 (0.9%)</td>
<td>0.46</td>
<td>3/42</td>
<td>2.6</td>
<td>Rowan and Downes [53]</td>
<td>8</td>
</tr>
<tr>
<td></td>
<td>Brookfield C.P., S.A. (34° 21'S, 139° 30'E)</td>
<td>16.0</td>
<td>300 (0.7%)</td>
<td>0.41</td>
<td>3/42</td>
<td>2.3</td>
<td>Jessup [54]</td>
<td>7</td>
</tr>
<tr>
<td>South-eastern Australia (nutrient-rich soils∗ — black soil plains)</td>
<td></td>
<td></td>
<td></td>
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<td></td>
<td></td>
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<tr>
<td>Woodland</td>
<td>Horsham-Nhill, Vic. (37° 00'S, 141° 55'E)</td>
<td>15.0</td>
<td>440 (1.3%)</td>
<td>0.45</td>
<td>4/44</td>
<td>3.1</td>
<td>Connor [55]</td>
<td>12</td>
</tr>
<tr>
<td>South-western Australia (nutrient-poor soils∗ — degraded lateritic podsols)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Tall open-forest</td>
<td>Karri forest, W.A. (34° 12'S, 115° 06'E)</td>
<td>15.4</td>
<td>1182 (40.5%)</td>
<td>0.65</td>
<td>3/42</td>
<td>3.0</td>
<td>Smith [56], Beard [57], Weston [58]</td>
<td>9</td>
</tr>
<tr>
<td>Open-forest</td>
<td>Perup River, W.A. (34° 20'S, 116° 27'E)</td>
<td>14.4</td>
<td>1055 (22.5%)</td>
<td>0.58</td>
<td>2/41</td>
<td>2.9</td>
<td>Williams [62], Smith [56], Beard [57]</td>
<td>8</td>
</tr>
<tr>
<td>Woodland</td>
<td>Dongolocking, W.A. (33° 04'S, 117° 14'E)</td>
<td>15.6</td>
<td>508 (1.8%)</td>
<td>0.55</td>
<td>2/41</td>
<td>2.4</td>
<td>Chapman et al. [63]</td>
<td>7</td>
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<tr>
<td>Low woodland</td>
<td>Coastal limestone, W.A. (32-33°S, 115° 45'E)</td>
<td>17.5</td>
<td>800 (7.0%)</td>
<td>0.50</td>
<td>2/40</td>
<td>2.0</td>
<td>Beard [57]</td>
<td>6</td>
</tr>
<tr>
<td>Woodland to shrubland</td>
<td>Woodline Hills, W.A. (28-30°S, 121-122°E)</td>
<td>18.5</td>
<td>&gt;300 (0.8%)</td>
<td>0.38</td>
<td>2/39</td>
<td>1.2</td>
<td>Beard and Webb [64], Dell et al. [65]</td>
<td>2</td>
</tr>
</tbody>
</table>

*nutrient-rich soils— black soil plains

*nutrient-poor soils— degraded lateritic podsols
<table>
<thead>
<tr>
<th>Plant formation</th>
<th>Locality</th>
<th>Mean annual temperature (°C)</th>
<th>Annual precipitation (mm)(runoff†%)</th>
<th>Evaporative coefficient (k) $\times 10^{-2}$ (per mm H$_2$O)</th>
<th>Overstorey &amp; understorey spp. per hectare</th>
<th>Annual shoot growth (t ha$^{-1}$)</th>
<th>Vegetation survey</th>
<th>Amphibia (no. of spp. per plant formation)</th>
<th>Amphibian survey</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Shrubland</strong></td>
<td>Durokoppin N.R., W.A. (31°25'S, 117°45'E)</td>
<td>17.5</td>
<td>339 (0.8%)</td>
<td>0.36</td>
<td>2/38</td>
<td>1.2</td>
<td>Muir et al. [66]</td>
<td>2</td>
<td>Daze [60], Tyler et al. [61]</td>
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<td></td>
<td>Yuna East N.R., W.A. (28°25'S, 115°12'E)</td>
<td>19.5</td>
<td>340 (0.8%)</td>
<td>0.35</td>
<td>2/38</td>
<td>1.0</td>
<td>Dell et al. [67]</td>
<td>1</td>
<td>Daze [60], Tyler et al. [61]</td>
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<td></td>
<td>McDermid Rock, W.A. (32°01'S, 120°44'E)</td>
<td>18.0</td>
<td>308 (0.7%)</td>
<td>0.34</td>
<td>1/38</td>
<td>1.0</td>
<td>Monk et al. [68]</td>
<td>1</td>
<td>Dell et al. [60], Tyler et al. [61]</td>
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<tr>
<td><strong>Heathland</strong></td>
<td>Southern Cross, W.A. (31°38'S, 119°30'E)</td>
<td>18.5</td>
<td>320 (0.7%)</td>
<td>0.34</td>
<td>0/50</td>
<td>1.0</td>
<td>George et al. [69]</td>
<td>1</td>
<td>Tyler et al. [61]</td>
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<tr>
<td><strong>Open-forest to woodland</strong></td>
<td>Belair N.P., S. A. (35°00'S, 138°40'E)</td>
<td>15.3</td>
<td>751 (5.6%)</td>
<td>0.55</td>
<td>2/41</td>
<td>1.8</td>
<td>Specht and Perry [70], [71]</td>
<td>6</td>
<td>Tyler [52]</td>
</tr>
<tr>
<td><strong>Open-forest to open-scrub</strong></td>
<td>Waterfall Gully in Cleland C.P., S.A. (35°00'S, 138°40'E)</td>
<td>12.9</td>
<td>970 (15%)</td>
<td>0.59</td>
<td>2/40</td>
<td>1.8</td>
<td>Specht and Perry [70], [71]</td>
<td>5</td>
<td>Tyler [52]</td>
</tr>
<tr>
<td></td>
<td>Flinders Chase N.P., Kangaroo Is, S.A. (35°52'S, 136°45'E)</td>
<td>14.5</td>
<td>635 (3.3%)</td>
<td>0.65</td>
<td>2/41</td>
<td>2.0</td>
<td>Wood [72], Baldwin and Crocker [73], Davies [74]</td>
<td>6</td>
<td>Tyler [52]</td>
</tr>
<tr>
<td><strong>Open-scrub</strong></td>
<td>Cox's Scrub C.P., S.A. (35°20'S, 138°45'E)</td>
<td>15.2</td>
<td>678 (4.0%)</td>
<td>0.56</td>
<td>2/40</td>
<td>1.5</td>
<td>Davies [74]</td>
<td>3</td>
<td>Tyler [52]</td>
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<tr>
<td></td>
<td>Innes N.P., S.A. (35°15'S, 136°55'E)</td>
<td>16.0</td>
<td>435 (1.3%)</td>
<td>0.50</td>
<td>2/39</td>
<td>1.3</td>
<td>Specht [27], Foale [75], Davies [74]</td>
<td>2</td>
<td>Wombey in Catling [43]</td>
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<tr>
<td><strong>Heathland</strong></td>
<td>Dark Island, S.A. (36°02'S, 140°29'E)</td>
<td>14.9</td>
<td>457 (1.3%)</td>
<td>0.42</td>
<td>0/45</td>
<td>1.0</td>
<td>Specht and Rayson [76]</td>
<td>1</td>
<td>Tyler [52]</td>
</tr>
<tr>
<td><strong>Wetlands to woodland</strong></td>
<td>Magela Ck, Jabiru, N.T. (12°52'S, 133°08'E)</td>
<td>18.0</td>
<td>1480 (&gt;60%)</td>
<td>0.50</td>
<td>23/35</td>
<td>7.2</td>
<td>Specht [77], Wilson et al. [78]</td>
<td>24</td>
<td>Tyler et al. [79]</td>
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<tr>
<td></td>
<td>Katherine Gorge, N.T. (14°28'S, 132°16'E)</td>
<td>26.9</td>
<td>959 (14.5%)</td>
<td>0.38</td>
<td>14/25</td>
<td>5.1</td>
<td>Christian and Stewart [80], Wilson et al. [78]</td>
<td>18</td>
<td>Tyler and Davies [81]</td>
</tr>
<tr>
<td><strong>Tussock grassland</strong></td>
<td>Barkly Tableland, N.T. (17-21°S, 135-138°E)</td>
<td>25.0</td>
<td>365 (0.9%)</td>
<td>0.30</td>
<td>10/20</td>
<td>3.8</td>
<td>Perry and Christian [82], Wilson et al. [78]</td>
<td>10</td>
<td>Tyler et al. [83]</td>
</tr>
<tr>
<td><strong>Wetlands to woodland</strong></td>
<td>Kununurra, W.A. (15°39'S, 128°42'E)</td>
<td>29.1</td>
<td>800 (7.0%)</td>
<td>0.45</td>
<td>19/30</td>
<td>6.3</td>
<td>Speck [84], Perry [85], Beard [86]</td>
<td>24</td>
<td>Tyler and Davies [81]</td>
</tr>
</tbody>
</table>
Table 1: Continued.

<table>
<thead>
<tr>
<th>Plant formation</th>
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<th>Amphibia survey (no. of spp. per plant formation)</th>
<th>Amphibian survey</th>
</tr>
</thead>
<tbody>
<tr>
<td>Open-forest</td>
<td>Groote Eylandt, N.T. (13° 49'S, 136° 38'E)</td>
<td>c. 26</td>
<td>1274 (almost nil)</td>
<td>0.40</td>
<td>12/22</td>
<td>4.2</td>
<td>Specht [77], Wilson et al. [78]</td>
<td>13</td>
<td>Tyler et al. [87]</td>
</tr>
<tr>
<td>Open-forest</td>
<td>Cobourg Peninsula, N.T. (11° 19'S, 132° 06'E)</td>
<td>c. 28</td>
<td>1331 (almost nil)</td>
<td>0.40</td>
<td>12/22</td>
<td>4.2</td>
<td>Specht [77], Wilson et al. [78]</td>
<td>13</td>
<td>Tyler et al. [87]</td>
</tr>
<tr>
<td>Open-forest</td>
<td>Melville Island, N.T. (11° 35'S, 131° 10'E)</td>
<td>c. 28</td>
<td>1466 (almost nil)</td>
<td>0.42</td>
<td>13/23</td>
<td>4.5</td>
<td>Wilson [88], Wilson et al. [78]</td>
<td>14</td>
<td>Tyler et al. [87]</td>
</tr>
<tr>
<td>Open-forest</td>
<td>Darwin, N.T. (12° 24'S, 130° 48'E)</td>
<td>28.0</td>
<td>1535 (almost nil)</td>
<td>0.45</td>
<td>14/25</td>
<td>4.9</td>
<td>Christian and Stewart [80], Wilson et al. [78]</td>
<td>17</td>
<td>Tyler and Davies [81]</td>
</tr>
<tr>
<td>Open-woodland</td>
<td>Tennant Creek, N.T. (19° 38'S, 134° 11'E)</td>
<td>25.2</td>
<td>362 (0.9%)</td>
<td>0.30</td>
<td>8/16</td>
<td>3.0</td>
<td>Perry and Christian [82], Wilson et al. [78]</td>
<td>7</td>
<td>Tyler and Davies [81]</td>
</tr>
<tr>
<td>Wetlands to woodland</td>
<td>Derby, W.A. (17° 18'S, 123° 38'E)</td>
<td>27.6</td>
<td>621 (3.1%)</td>
<td>0.39</td>
<td>11/21</td>
<td>4.1</td>
<td>Speck and Lazarides [89], Beard [86]</td>
<td>14</td>
<td>Tyler et al. [90]</td>
</tr>
<tr>
<td>Wetlands to woodland</td>
<td>Gibb River, W.A. (16° 26'S, 126° 26'E)</td>
<td>c. 28</td>
<td>764 (5.9%)</td>
<td>0.38</td>
<td>11/21</td>
<td>4.0</td>
<td>Speck and Lazarides [89], Beard [86]</td>
<td>13</td>
<td>Tyler et al. [90]</td>
</tr>
</tbody>
</table>

**Northern Australia (medium-nutrient soils\(^{\ast}\)—deep lateritic earths)**

\(^{\ast}\) Relative to plant formations on medium-nutrient soils (Figure 4), annual shoot growth (dry weight measured in t ha\(^{-1}\)) in major plant formations on nutrient-poor soils is about 60% less, 50% less on extremely nutrient-poor soils, and 130% more on nutrient-rich soils [16, 36].

\(^{\dagger}\) Annual stream flows from the major drainage basins in Australia have been collated by the Australian Water Resources Council [92]. These estimates of runoff from the landscape show an exponential relationship to annual precipitation [16].

\[
\text{Runoff/Precipitation} = 0.175 + 10^{-0.002 \times \text{Precipitation}} \quad (r = 11, r^2 = 0.60).
\]

Minimal runoff occurs from the deep lateritic earths of northern Australia (see the hydrological survey of the Weipa area in North Queensland, reported by Specht et al. in [93]).

---

Table 1: Continued.

<table>
<thead>
<tr>
<th>Plant formation</th>
<th>Locality</th>
<th>Mean annual temperature (°C)</th>
<th>Annual precipitation (mm) (runoff(^{\dagger}))</th>
<th>Evaporative coefficient (k) (\times 10^{-2}) (per mm H(_2)O)</th>
<th>Overstorey &amp; understorey spp. per hectare</th>
<th>Annual shoot growth (t ha(^{-1}))</th>
<th>Vegetation survey</th>
<th>Amphibia survey (no. of spp. per plant formation)</th>
<th>Amphibian survey</th>
</tr>
</thead>
<tbody>
<tr>
<td>Tall shrubland</td>
<td>Alice Springs, N.T. (23° 49'S, 133° 53'E)</td>
<td>20.6</td>
<td>255 (0.6%)</td>
<td>0.33</td>
<td>5/12</td>
<td>2.1</td>
<td>Speck and Lazarides [89], Beard [86]</td>
<td>5</td>
<td>Tyler and Davies [81]</td>
</tr>
</tbody>
</table>
3. Results

3.1. Evaporative Coefficients (k) of Sampling Sites. The evaporative coefficients computed for monsoonal northern Australia (Table 1) indicate that the sampling regions range in climate from the subhumid \((k = 0.55 - 0.45 \times 10^{-2})\) to the semiarid zone \((k = 0.45 - 0.35 \times 10^{-2})\).

In contrast, the evaporative coefficients computed for temperate southern Australia (Table 1) indicate that the sampling regions range in climate from the humid \((k = 0.75 - 0.55 \times 10^{-2})\) to the semiarid zone \((k = 0.45 - 0.35 \times 10^{-2})\).

3.2. Annual Foliage Shoot Growth (ASG, \(t\ ha^{-1}\)) in Overstorey Strata. Solar energy fixation by photosynthesis in the leaves of both overstorey and understorey plants within a plant community is influenced by the horizontal (FPC) and vertical (LAI) distribution of foliage, ambient temperature, and high-energy phosphorus and nitrogen compounds in the mature leaves [9]. So far, it has been only possible to assess the annual biomass production of new foliage leaves per hectare of plant communities—from the temperate south to the tropical north on medium-nutrient soils in eastern Australia (Equations (2)–(4)). This nutrient class—defined by Kulczynski Symmetric analyses of the major foliar nutrients in overstorey and understorey leaves [16]—is characteristic of the clayey podsolic soils in temperate Australia and the deep lateritic earths in monsoonal Australia (Table 1).

The responses of a range of eucalypt and heath species to the application of phosphate to soils (in the glasshouse and in the field) have been assessed (Figure 3(a)). The nutrient-poor lateritic soils characteristic of much of south-western Australia reduce the annual growth of foliage shoots (per hectare) by roughly 60%; extremely nutrient-poor sandy soils show a reduction of about 50%. The fertility of the nutrient-rich black soils of the coastal plains of northern Australia and in the inland cracking clays of the Barkly Tableland, the Darling Downs, as well as the Riverina and the Wimmera-Tatiara Districts increases annual growth by 120%–130% above that of the medium-nutrient soils. Adjusted values of the annual growth of foliage shoots (ASG, \(t\ ha^{-1}\)) are tabulated in Table 1.

3.3. Species Richness of Vascular Plants and Annual Foliage Shoot Growth. The species richness (number of species per hectare) of vascular plants in the overstorey stratum of major plant communities growing on medium-nutrient soils in the arid to the humid climatic zones of tropical, subtropical,
Figure 4: (a) The relationship between species richness ($N$, number of species per hectare) of overstorey species in major plant communities—growing on soils of medium nutrient status—with Annual Shoot Growth (biomass per hectare) in tropical, subtropical and temperate eastern Australia (after [14, 15]). Overstorey $N = 1.95 \times 10^{0.157 \text{ASG}}$ ($n = 15; r^2 = 0.98$). (b) The relationship between species richness ($N$, number of species per hectare) of overstorey and understorey species in major plant communities—growing on nutrient-rich to nutrient-poor soils—with Annual Shoot Growth (ASG, biomass per hectare) in tropical and temperate Australia (Table 1). Temperate overstorey $N = 0.99 \text{ASG} + 0.35$ ($n = 20; r^2 = 0.94$). Temperate understorey $N = 1.93 \text{ASG} + 36.5$ ($n = 20; r^2 = 0.93$). Tropical overstorey $N = 3.33 \text{ASG} - 2.13$ ($n = 13; r^2 = 0.98$). Tropical understorey $N = 4.30 \text{ASG} + 3.61$ ($n = 13; r^2 = 0.99$).

and temperate eastern Australia is related exponentially to annual shoot growth (ASG, t ha$^{-1}$) in the overstorey stratum (Figure 4(a), after [16]).

As all the sampling sites in tropical northern Australia lie within the subhumid to semiarid climatic zone, the species richness data assembled in Table 1 fall on the essentially linear part, near the origin, of the exponential curve (Figure 4(a)). This linear relationship between species richness of vascular plants and annual shoot growth is shown in Figure 4(b).

Even though the sampling sites in temperate Australia extend from the semiarid to the wetter humid climatic zone, these data are located on the linear part, near the origin, of the exponential curve (Figures 4(a) and 4(b)).

3.4. Species Richness of Vascular Plants and Amphibia. The species richness of vascular plants in the overstorey and understorey strata of major Floristic Groups—also the Amphibia in wetlands associated with these Floristic Groups—is tabulated in Table 1. Even though the climate of most of the Top End of the Northern Territory is largely subhumid (and only rarely in the humid zone), the species richness of the overstorey strata reaches 23 species per hectare, markedly in contrast to the species richness (six per hectare) of the overstorey in the humid climatic zone of temperate Australia. The species richness of the understorey, however, is lower in the monsoonal north (12 to a maximum of 35 species per hectare) in comparison to that in temperate Australia (39–46 species per hectare). The number of amphibian species recorded (Table 1) in temperate Australia (one to 17 species) is less than the numbers recorded in the monsoonal north (five to 24 species).

The number of amphibian species is strongly correlated with the number of species of vascular plants—overstorey and understorey strata—in the major Floristic Group of each region (Figure 5).

3.5. Species Richness of Amphibia and Annual Foliage Shoot Growth. Species richness (per region) of Amphibia declines continuously in the macroclimatic gradient from the humid to the semiarid zone in south-eastern Australia, south-western Australia, and the wet-dry tropics of northern Australia (Figure 6(a)). On the medium-rich soils of south-eastern Australia, species richness of Amphibia is higher, by three to five species, than that of species recorded on the nutrient-poor lateritic soils of south-western Australia. Species richness is further reduced on the very infertile sandy soils of the Mount Lofty Ranges, southern Yorke Peninsula, and Kangaroo Island in South Australia. Although only one area of nutrient-rich soils—the Wimmera District of Victoria—is located in southern Australia, the species
4. Discussion

4.1. Annual Foliage Shoot Growth (ASG, t ha\(^{-1}\)) in Overstorey Strata. The leaves of all evergreen overstorey and understorey plants in open-structured plant communities experience considerable stress during the long dry seasons experienced throughout Australia—during the winter months in the monsoonal north and during the summer months in the Mediterranean-type climate in the south. The leaves of C4-photosynthetic, perennial grasses that dominate the savanna (grassy) understorey may appear dead but show a remarkable power of cytoplasmic resurrection as soon as rains fall [101, 102].

The development and survival of the multitude of new leaves, which are produced annually in overstorey and understorey strata during the short periods of foliage growth in all life forms, are affected by turbulent aerodynamic fluxes (frictional, thermal, evaporative) in the atmosphere as it flows over and through a plant community.

Frictional forces “abrade” developing shoots at the edges of canopies of all component life forms in a plant community so that \(\sum (\text{FPC}_o + \text{FPU}_o)\), the sum of the foliage projective covers of the overstorey (FPC\(_o\)) and understorey (FPC\(_u\)) strata, is a constant in space and time in both dryland and wetland ecosystems [6, 16, 40].

Evaporative fluxes in the turbulent atmosphere above and within a plant community influence the movement of water in the transpiration stream that contains nitrogen and phosphorus ions. These nutrients are essential for the formation of high-energy compounds in mature leaves, from whence they are translocated to developing shoot apices to form the cytoplasmic contents and cell walls of new leaves. As “drag” (shearing stress) in the boundary layers above a plant community is minimal [103], the balance between cell wall to cytoplasm per unit area of mature leaf—leaf specific weight (LSW)—is constant in all overstorey and understorey strata [6, 16]. It follows that stomatal resistances and carbon isotope ratios in the leaves of these strata will be representative of aerodynamic conditions during the seasonal periods of foliage shoot growth in each stratum.

Within the horizontal constraints of overstorey and understorey, defined by \(\sum (\text{FPC}_o + \text{FPU}_o)\), the number of leaves produced vertically on developing foliage shoots is dependent on the flow of water and nutrients in the transpiration stream during the short period of shoot growth each year; a gradient in shoot elongation is apparent in overstorey trees as available water increases [6]. Ambient temperature during the period of active shoot growth also influences the number of leaves produced on foliage shoots in the 10\(^\circ\)C gradient from temperate to tropical Australia. With more leaves per foliage shoot, the photosynthetic potential of the plant community is enhanced and results in increased annual growth (biomass per hectare) and determines the ultimate height of overstorey stratum in a plant community [11, 12, 16].

The higher temperatures in northern Australia increase annual solar energy fixation by foliage in the overstorey stratum. As the main growth phenophases of understorey plants in northern Australia [104, 105] occur when solar
radiation is essentially vertical [12, 16, 106], high ambient temperatures in this stratum result in marked higher stem-root respiration and thus affect the primary production of nonphotosynthetic organs [11, 12]. In southern Australia where ambient temperatures are 10°C cooler than in the north and the sun’s rays are oblique, annual growth (ASG, t ha⁻¹) of foliage shoots in both overstorey and understorey strata is reduced; however, as leaf respiration is lower in the cooler south than in the tropical north of Australia, net photosynthesis of each stratum is higher throughout the year [11, 12].

4.2. Species Richness of Vascular Plants and Annual Foliage Shoot Growth. In combination with aerodynamic fluxes, available soil water and soil nutrient status determine the fixation of solar energy as biomass (expressed as dry weight) of foliage shoots produced annually in every hectare of vegetation [11, 16, 107, 108]. Species richness (number of species per hectare) and stem density (number of stems per hectare) of the overstorey of a plant community on medium-nutrient soils are exponentially correlated with annual shoot growth (ASG, t ha⁻¹) (Figure 4(a), after [11, 12, 14–16, 18]).

In this paper, the species richness (number of species per hectare) of vascular plants in the nutrient-rich, medium-nutrient, nutrient-poor, and extremely nutrient-poor ecosystems (Figure 3(a)) of southern temperate and northern tropical Australia is plotted against annual shoot growth along the linear section, near the origin, of the exponential curve correlating species richness from the semiarid to the perhumid climatic zone in tropical, subtropical, and temperate Australia (Figure 4). These linear regressions indicate that the species richness of overstorey plants is higher in the tropics (5–23) than in temperate southern Australia (2–6); the inverse is found in understorey plants with 40–46 in temperate Australia and 12–35 species per hectare in monsoonal northern Australia.

The species richness of understorey plants in dryland plant communities is inversely related to the degree of
shading produced by the foliage projective cover of the overstorey stratum [5–7, 40].

In wetland plant communities, however, species richness of overstorey and understorey strata both decreases as the number of months of waterlogging increases [7]. Nevertheless, the sum of foliage projective covers in the overstorey and understorey strata—\(\sum(FPC_o + FPU_u)\)—remains relatively constant along the wetland continuum [7].

Ambient temperature during the period of active shoot growth affects cell formation in developing leaves of the overstorey strata in the 10°C gradient from temperate to tropical Australia; increased species richness of overstorey strata (Figure 4(b)) appears to be associated with increased annual shoot growth (ASG, t ha\(^{-1}\)). On the other hand, the higher temperatures associated with almost vertical solar radiation in the north [12, 16, 106], in contrast to the oblique, low-intensity rays in the south, result in higher leaf respiration (and, thus, reduced net photosynthesis) of this stratum throughout the year; a reduction in the species richness in the understorey stratum results (Figure 4(b)).

4.3. Species Richness of Amphibia and Annual Foliage Shoot Growth. Within each major plant community, the species richness (per plant community) of resident consumers (small mammals, birds, lizards, elapid snakes, and also Amphibia in associated wetlands) is also correlated with annual shoot growth [16, 17, 23, 24]. As the different trophic levels of consumers are dependent on the energy fixed by the plant community, it follows that the species richness of these faunal groups will follow the species richness of plants in each ecosystem (Figure 5).

Most of the above surveys of the species richness of vertebrates associated with major plant communities were made in southern Australia on medium-nutrient soils; only the survey of small mammals by Braithwaite et al. [21] was made in the north. Pianka and Schall [109] did assemble data on the species richness of marsupials in grid squares with sides about 240 km long throughout Australia; their data are linearly correlated with the number of major plant communities recorded in each of their grid squares [16, 39].

This paper examines the species richness of Amphibia in both northern and southern Australia on a range of ecosystems from extremely nutrient-poor to nutrient-rich habitats (Figure 3(a)). The species richness of Amphibia shown by the five distinct linear regression lines of Figure 6(a) (plus that of the nutrient-rich Wimmera Region) coalesces when the data are plotted against the annual growth of foliage shoots (ASG, t ha\(^{-1}\)) of the major plant communities at each sampling site (Figure 6(b)).

4.4. Maintenance of Species Richness of Native Vascular Plants and Vertebrates. Over the last fifty million years, the original Gondwanan flora (rainforest flora ± primitive eucalypts, temperate Nothofagus rainforest, savanna grasses, heathy plants, desert chenopods, etc.) has survived on the Australasian Tectonic Plate as it drifted northward from the subtropical Antarctic continent [16, 19, 110, 111]. The gradual development of the Southern Ocean between Antarctica and Australia induced an arid climate over much of the continent [112]; rainforests became confined to well-watered habitats, being replaced by open-structured vegetation across the continent.

The fixation of solar energy to produce foliage shoots (per hectare) in a plant community is obviously dependent on the macroclimate and soil fertility of the site. Any environmental change will affect the annual production of foliage shoots (per hectare) and, over a long period of time, result in a change in species richness of producers, consumers, and decomposers in the ecosystem [16, 113–117]. The reversal of ammonium production from decomposing litter in native ecosystems to that of nitrate production after disturbance [13, 118] enables the invasion of thin-leaved pioneers and introduced plants [119], further upsetting species richness of flora and fauna within an ecosystem [117].

As well, phosphate pollution on these soils has been demonstrated as a threat to the survival of much of the Australian heathland, savanna, and even rainforest vegetation and leads to the invasion of many exotic plants in the gaps that result [16, 24, 36, 38, 115–117, 120]. Nutrient transport, by run-off and seepage, into wetlands from undisturbed terrestrial plant communities, is minimal, especially as the majority of Australian soils contain a high percentage of kaolinite that fixes phosphate and molybdate ions within their clay lattice. Phosphate is conserved from leaching—as polyphosphate granules in roots and associated rhizosphere organisms when it is released as orthophosphate during the decomposition of leaf litter—in most of the Gondwanan-derived ecosystems [121–124]. Phosphate pollution in disturbed dry-land ecosystems surrounding a wetland, however, will accumulate in the surface organic layers of the soil [38] whence it can leak into drainage water [125] and affect the structure, growth, and biodiversity of wetland vegetation [126].

Changes in species richness of both plants and animals within most Australian ecosystems in conservation reserves [19, 127] will result if environmental perturbations affect the annual biomass production of a plant community. Global warming will increase the evaporative power of the atmosphere world-wide, independently of any change in rainfall; both \(\sum(FPC_o + FPU_u)\) of the understorey and overstorey foliage and the ecomorphological attributes of leaves will be altered and influence primary productivity [16, 113, 128]. With a 2°C rise in global temperature, the structure of Australian plant communities, from the tropical north to the temperate south, will change: tall open-forest → open-forest → woodland → open-scrub → tall shrubland → low shrubland → desert.

Phosphate and nitrate pollution will induce marked changes in the Australian vegetation—savanna, heathland, and rainforest ecosystems—that has evolved on the nutrient-poor soils that covered much of Gondwanaland [16, 36, 38, 115–117, 119].

An increase in the populations of soil chytrid fungi will be promoted by phosphate pollution (Figure 3(b), after [37]) and may pose a threat to Amphibia [129]. A total of 27 species belonging to the aquatic fungal group Chytridiales—all involved in the decomposition of cellulose, chitin, and keratin—were recorded in soil samples from
southern Australia, from Perth to Wilson’s Promontory, and as far inland as Alice Springs [33, 34, 37, 121, 130]. The easily cultured chytrid, *Rhizophlyctis rosea*, showed increased activity in soil samples whenever phosphate had been added as a fertiliser (Figure 3(b)). It must be expected that the input of phosphate pollution into wetlands from urban and rural communities will induce an increase in the activity of all chytrid species and thus may impose a threat to the survival of many individual frogs currently living in these habitats.

Exotic grazing animals (domestic animals, rabbits, etc.) prefer the thin-leaved plants that have occupied any disturbed site within the thick-leaved native plants that have survived and evolved in Australia during the last 100 million years [110]. Even the developing leaves of overstorey eucalypts and understorey shrubs, in habitats where the supply of high-energy phosphorus and nitrogen compounds is above normal, are more attractive to attack by small insects and mites [6, 131, 132]—thus leading to increased allergic faecal dust in the atmosphere [16, 133, 134].

Atmospheric salinity, blown inland by off-shore winds, will further affect annual shoot growth of C3-photosynthetic overstorey species and thus lead to a marked reduction in species richness [16, 135, 136]. Trace amounts of sodium, however, are essential for the C4-photosynthetic, perennial grasses and chenopods that dominate the savanna (grassy) understorey of tropical, subtropical, and even temperate Australia and in the vast Arid Zone of the continent [137, 138].

It is thus apparent that the species richness of the flora and fauna of the Gondwanan heritage of Australia is under threat from phosphate pollution, atmospheric salinity, and Global Warming. The first perturbation may lead to increased annual shoot growth (per hectare) and the last two perturbations to a reduction in annual shoot growth. Increased annual shoot growth of native plant communities leads to the invasion of exotic plants and animals; decreased annual shoot growth results in a reduction in species richness per hectare. The effect of these two perturbations on the ultimate composition and species richness of many Australian ecosystems needs intensive investigation.

An understanding of the physiochemical processes that operate within an entire ecosystem to maintain the species richness of the plant community (and associated consumers and decomposers) is essential for their scientific management.

**Acknowledgments**

The data presented here have been accumulated over many years as a result of research funding from several agencies. Mike Tyler would particularly like to acknowledge the support of the Australian Research Committee and the Utah Foundation. Information on the floristics of overstorey and understorey strata in plant communities has been published in ecological surveys made of every 30° latitude × 30° longitude grid cell throughout Australia. These data have been collated in sixteen large data banks and analysed by the classificatory program TWINSPLAN [39]—with support from the Australian Heritage Commission.

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