The Promotion Effect and Mechanism of Methanoic Acid on the Photoelectrocatalytic Degradation of Fulvic Acid

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1. Introduction

Fulvic acid (FA) is a primary kind of natural organic matters (NOMs) in the water body [1, 2] and it is also a main precursor of disinfection by-products (DBPs), which is known as a kind of harmful materials, leading to carcinogenic, teratogenic, and mutagenic effects [3]. Thus, removing FA from water is urgently needed. However, due to refractory nature of FA, only limited methods can be applied such as granular activated carbon adsorption, enhanced coagulation, and membrane filtration. Among them, even if the enhanced coagulation shows its advantage to restrain large quantity of NOMs including colloid humic acid, it is still ineffective to the soluble FA [4, 5].

Recently, the advanced oxidation technologies have been a research’s hot spot in removal of NOMs like FA, because it provides hydroxyl radical with strong oxidation ability to decompose organics into small harmless molecules such as CO₂ and H₂O [6–8]. In particular, the photoelectrocatalytic (PEC) technology attracts much more attention due to the remarkable merits, such as low cost, good oxidation ability, high efficiency, and no secondary pollutant [9–11]. However, the PEC technology has been seldom reported to be used in the degradation of FA [12–15].

TiO₂ is a semiconductor material, which is widely used as a photoanode in the PEC technology and its excellent performance has been tremendously reported [16–21]. In this study, TiO₂ nanotube arrays (TNAs) are chosen as the photoanode material, because they have larger specific surface area, more stable structure, and better PEC properties than the other forms of TiO₂. In particular, it could degrade more varieties of refractory organic matter [22–24]. On the other hand, the degradation performance of organics could be influenced by many factors, involving not only the photocatalyst but also the reaction medium. For instance, some chemical matters, such as phosphate, carbonate, and tertiary butyl alcohol, could restrict the activity of hydroxyl radical and further repress the PEC degradation efficiency of organics [25, 26].

In the limited previous researches, the feasibility of photocatalytic technology used in the degradation of FA is proved [12–15]. However, the degradation of FA is inefficient and time-consuming, or the preparation of the catalytic material is very complex [13, 14]. In this study, we proposed...
2. Materials and Methods

2.1. Materials. All the reagents in this study were analytically pure and purchased from Sinopharm Chemical Reagent Company (Shanghai, China). All solutions were mixed with high-purity deionized water (18 MΩ) purified by a Milli-Q purification system (Millipore Corporation, Billerica, MA). The NaNO₃ solution served as a supporting electrolyte in the samples.

2.2. Preparation of the TNA Electrode. The TiO₂ nanotube electrodes were prepared by the electrochemical anodic oxidation method, which is reported in our previous work [28, 29]. In a typical process, the TiO₂ nanotube arrays were grown after the potentiostatic anodization at 20 V using a platinum cathode in a solution of 1 mol L⁻¹ NaHSO₄, 1 mol L⁻¹ NaF, and 0.2 mol L⁻¹ trisodium citrate for 6 h. Then as-prepared samples were annealed in a laboratory muffle furnace at 500 °C for 3 h to form anatase TiO₂ nanotube arrays (TNAs).

2.3. Characterization of TNA Electrode. The phase of TNA electrode was identified by X-ray diffraction (XRD-6100, Shimadzu, Japan), using Cu KR (λ > 0.15406 nm) radiation at 40 kV and 30 mA at a scanning rate of 10°/min in the 2θ range from 10° to 90° at room temperature. The surface morphology of TNA electrode was investigated by field emission scanning electron microscopy (Nova NanoSEM 450, FEI Company, USA) under a voltage of 5 kV.

2.4. Apparatus and Methods. A thin-layer equipment shown in Figure 1 served as the reactor, where the PEC degradation process was carried out. The thin-layer reactor consists of the following six major sections: the TNAs working electrode, the saturated Ag/AgCl reference electrode, the platinum counter electrode, the flow inlet, the flow outlet, and a quartz window with a diameter of 1 cm. The reaction cell in the reactor is built by two coherent polytetrafluoroethylene planks. The cell’s thickness was controlled no more than 0.1 mm in order to reduce the volume of the solution, shorten the time of degradation and the distance of the charge transfer from the solution to the surface of electrode, and ensure the light transmittance in the meanwhile. A 365 nm ultraviolet LED is chosen to be the light resource. The thin-layer reactor was controlled and monitored by an electrochemical workstation (CHI 610D, Chenghua, Shanghai), which was connected to a computer to record the photocurrent response signals of the PEC degradation.

The degradation rate (α) of organics is determined based on an exhaustive degradation model [27]. In this model, the organics in the thin-cell reactor are rapidly photoelectrocatalytically oxidized and a set of photocurrent-time profiles was then obtained from exhaustive degradation of organics. As shown in Figure 2, photocurrent observed is the sum of two components, one from the oxidation of organics and the other from the oxidation of water (I_{background}). The decay of I_{total} to the I_{background} level implies that the organics have been partly oxidized and generate the corresponding net charge quantity.
However, under the given condition, the oxidized proportion $\alpha$ should be a constant assuming that the organic compounds could be indiscriminately oxidized. It is possible to obtain $\alpha$ from the ratio of the captured $Q_{net}$ and theoretical value of the net charge quantity (Th$Q_{net}$) of the whole sample, which is well matched with net charge quantity measured by the exhausted oxidation of the whole sample ($Q_{net}$) according to

$$\alpha = \frac{Q_{net}}{ThQ_{net}}.$$  \hspace{1cm} (1)

This method could facilitate the measurement of the degradation rate of different organics according to the amount of electrons.

2.5. Degradation Experiments. 25–100 mg/L FA and 8–60 mg/L MA were used as a model pollutant. The PEC experiments were performed in the thin-layer PEC reactors. The supplied voltage, electrolyte concentration, and UV light intensity was controlled in a certain range. The photocatalysis (PC) experiment was performed by using the same system without applying an external potential and the electrolysis (EC) experiment was performed without UV illumination.

3. Results and Discussion

3.1. Characterization of Photoanode. The FE-SEM images of the TNAs electrode are shown in Figures 3(a) and 3(b). As can be seen, the nanotubes are highly ordered and well aligned, with the diameter of ~100 nm and wall thickness of ~10 nm [28]. The cross-section image shows that the nanotubes are ~1.5 $\mu$m in length and closely combined to the barrier layer. The XRD patterns of the as-annealed TNAs electrode are shown in Figure 3(c). The nanotube without calcinations maintained an amorphous structure in Figure 3(c), B, while the amorphous regions were gradually crystallized to form anatase phases via the calcinations at 500°C in Figure 3(c), A. It can be seen from pattern A that the sample possesses characteristic peaks at 25.35° (101), 38.1° (004), and 48.2° (200) for the anatase phase. The titanium peaks can also be found owing to the Bragg reflection of Ti substrate.
3.2. Parameters Affecting the PEC Degradation. The catalytic mode, bias voltage, light intensity, and the concentration of electrolyte are first investigated in order to optimize the PEC degradation. Take 25 mg/L FA as an example; various degradation processes were applied including photocatalysis (PC), electrolysis (EC), and photoelectrocatalysis (PEC) and the experimental results are summarized in Figure 4. As can be seen, the photocurrent values of both the PC and EC are much smaller than PEC, implying that both PC and EC mode are ineffective for the degradation of FA. The TNAs electrode has been approved as an excellent photoanode in the previous reports [27, 28] since it has an advantage of fast electron transfer. When the given electric bias was supplied to the photocatalyst, it promoted the separation of photogenerated holes and electron. This charge separation could prevent the photoinduced electron and the hole from recombining and consequently lead to higher degradation efficiency. In short, the bias voltage and the UV light illumination have a synergistic effect in the process of PEC degradation.

Table 1 shows the comparison of different bias voltages, light intensity, and electrolyte concentration applied in the PEC process. Based on the measuring method used in this study, $Q_{\text{net}}$ was directly used for analyzing the optimum degradation conditions. In terms of bias voltage, no obvious change of $Q_{\text{net}}$ is found in the PEC degradation under the bias voltages ranged from 1.0 V to 2.5 V. However, using a larger bias voltage like >3.0 V, a deviant increase of $Q_{\text{net}}$ appears possibly due to water electrolysis reaction. Thus, a bias voltage of 2 V is preferred.

In the process of PEC degradation, the light intensity could influence the speed of the reaction. However, it has no impact on $Q_{\text{net}}$ via the result shown in Table 1. Considering the fact that high light intensity could result in a fast and stable reaction, a light intensity of 37.4 mW/cm² was applied for the degradation experiment.

The concentration of supporting electrolyte has a direct relationship with the resistance of the solution that could influence the transportation of the photo electrons. As shown in Table 1, the value of $Q_{\text{net}}$ starts to keep stable with the concentration of NaNO₃ increasing to over 2.0 mol/L. Thus, 2.0 mol/L NaNO₃ solution is used as supporting electrolyte.

In conclusion, the optimum condition used in the following PEC experiment is applied at 2.0 mol/L NaNO₃ electrolyte, 2 V bias voltage, and 37.4 mW/cm² light intensity.

3.3. The PEC Degradation Property of FA and MA. Figure 5 shows the photocurrent-time profiles and degradation rate ($\alpha$) of (a), (b) FA and (c), (d) MA, respectively. As can be seen in Figure 5(a), the photocurrent gradually shifted to the right with the increase of FA concentration from 0 (baseline) to 100 mg/L. It indicates that a higher concentration of FA needs a longer degradation time and, accordingly, a larger $Q_{\text{net}}$. The degradation rate ($\alpha$) of FA can be calculated by (1) based on the photocurrent-time profiles in Figure 5(a). The result shows that $\alpha$ only ranges from 36% to 41%, indicating only small portion of FA degraded in the PEC process. It can be deduced that most of hydroxyl radicals have been consumed in the beginning of the reaction, while the residual FA and the intermediate products could not be continuously degraded, finally resulting in a lower degradation rate. Thus, we hypothesize that if more hydroxyl radicals were produced in the PEC process, the degradation rate of FA could be improved.

Compared to FA, MA shows a more degradable property. As shown in Figure 5(c), a larger photocurrent can be observed even in a smaller concentration of MA, implying more electrons generated from the degradation of MA. The degradation rate of MA is also calculated based on Figure 5(c). As can be seen from Figure 5(d), the $\alpha$ values of MA at different concentrations are close to 100%, indicating a complete degradation. The excellent degradability of MA is possibly due to its small molecule.

3.4. The Promotion Effect of MA on the Degradation of FA. The promotion effect is investigated by adding different concentrations of MA to 50 mg/L FA in the PEC degradation. In this study, 8, 15, 30, 45, and 60 mg/L of MA were used. Figure 6(a) shows the photocurrent-time profiles obtained from the PEC degradation of FA mixed with different concentrations of MA. The total $Q_{\text{net}}$ of the mixture was calculated, respectively. Because MA can be completely
degraded, $Q_{\text{net-MA}}$ should be equal to $T Q_{\text{net-MA}}$. Therefore, $Q_{\text{net-FA}}$ could be calculated by subtracting $T Q_{\text{net-MA}}$ from the total $Q_{\text{net}}$, and then the degradation rates ($\alpha$) of FA could be calculated according to (1). As shown in Figure 6(b), $\alpha$ increases from 40% to 100% with the increase of MA concentration from 0 to 60 mg/L. Obviously, MA has a promotion effect on the PEC degradation of FA. For example, when 45 mg/L MA was added, $\alpha$ of 50 mg/L FA reaches 100%. In other words, more than 45 mg/L additional MA could promote the complete degradation of 50 mg/L FA.

The mechanism was also discussed herein. As known, TBA is a researched hydroxyl radical scavenger [25, 26], which can reduce the concentration of hydroxyl radical. Take the mixture of 50 mg/L FA and 45 mg/L MA as an example; different concentrations of tert-butyl alcohol (TBA) were added to the mixture and the result was shown in Figure 7. As can be seen, $\alpha$ of the mixture obviously reduces with the increase of TBA concentration. As 30 mg/L TBA is added, $\alpha$ reduces about 10%. This is because the promotion effect of MA on the PEC degradation of FA is restrained when the amount of hydroxyl radial is limited. Thus, it could be deduced that an enormous amount of hydroxyl radicals is generated because of the excellent degradability of MA. The additional hydroxyl radicals could continuously degrade both FA and the intermediate products in the degradation reaction.

4. Conclusion

In this study, an obvious promotion effect of MA on the PEC degradation of FA was observed. The degradation rate of FA alone is only about 40%, while it could be significantly improved by addition of MA. With 45 mg/L MA mixed with 50 mg/L FA, the degradation rate of FA can achieve 100%. The
mechanism is also investigated where the enormous amount of hydroxyl radicals, which is generated in the degradation of MA, could maintain the continuous degradation of both FA and the intermediate products. This work provides a new way to promote the PEC degradation of FA and a direction to remove humus from polluted water.

**Competing Interests**

The authors declare that there are no competing interests regarding the publication of this paper.

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