Research Article

Synergistic Adsorption and Photocatalytic Activity under Visible Irradiation Using Ag-ZnO/GO Nanoparticles Derived at Low Temperature

Viet Ha Tran Thi,1 The Ha Cao,1,2 Tri Nhut Pham,3,4 Tien Thanh Pham,5 and Manh Cuong Le6

1Master’s Program in Environmental Engineering, Vietnam Japan University, Hanoi, Vietnam
2Center for Environmental Technology and Sustainable Development (CETASD), Hanoi, Vietnam
3Center of Excellence for Green Energy and Environmental Nanomaterials, Nguyen Tat Thanh University, Ho Chi Minh City, Vietnam
4NTT Hi-Tech Institute, Nguyen Tat Thanh University, Ho Chi Minh City, Vietnam
5Master’s Program in Nano Technology, Vietnam Japan University, Hanoi, Vietnam
6Department of Building Materials, National University of Civil Engineering, Hanoi, Vietnam

Correspondence should be addressed to Viet Ha Tran; ttv.ha@vju.ac.vn

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1. Introduction

Developing semiconductor photocatalysts has been considered a promising green technology for solving environmental issues. Among various semiconductors, zinc oxide (ZnO) has been considered one of the most promising photocatalysts because of its outstanding properties, such as physical and chemical stability, low cost, and nontoxicity. Thus, ZnO has been widely used for many different applications such as in optical materials, sensors, solar energy conversion devices, and photocatalysts for pollutant treatment. ZnO also has several advantages over TiO2—the most popular photocatalyst [1], which includes higher thermal stability and easier and less expensive synthesis [2]. However, the photocatalytic performance of ZnO is reduced because of its wide band gap (theoretical ~ 3.2 eV at room temperature) [3] and high recombination rate between photogenerated electrons from the conduction band (CB) and holes from the valence band (VB) [4]. Such a noble metal modification can also influence the surface properties, in particular, by introducing hydroxyl groups on the surface of the photocatalysts [5–7]. However, these modifications require calcination at high temperature. 

Ag-ZnO/graphene oxide (AG-ZnO/GO) nanocomposite was synthesized via facile aqueous solution reactions at low temperature in order to improve the photocatalytic activity for cationic dye removal under visible light irradiation. Analytical techniques were carried out in order to determine the abilities including structure, state of elements, morphology, and surface area of synthesized materials. Ag-ZnO/GO nanocomposite presented an extremely high removal rate of methylene blue (MB) not only under UV light (over 99% removal) but also under visible light (85% removal) during the same irradiation time. In this study, initial process parameters of catalyst dosage, MB concentration, and pH of the solution were also examined for MB removal efficiency effects. The proposed mechanisms for the increased removal of MB by Ag-ZnO/GO nanocomposite under visible irradiation include increased photocatalytic degradation, mainly due to increased charge transfer capacity by lowering band gap energy; minimized recombination of the excited electron-hole pairs of ZnO with the addition of Ag into the ZnO crystal lattice; and an increased adsorption capacity with the addition of GO with high surface area and semiconductor function with zero band gap energy.
Graphene sheets possess a unique two-dimensional layered structure of sp²-hybridized carbon atoms, and thus, they can exhibit novel electronic properties as zero band gap semiconductors. Ahmad et al. synthesized graphene-Ag/ZnO nanocomposite via a solvothermal method [8]. However, it is not easy to synthesize due to high levels of impurities with graphene; thus, oxidized forms are usually used as a substitute for graphene in many research studies. In addition, graphene oxide (GO) provides a large scaffold for anchoring various substances owing to its large specific surface area [9] and two-dimensional planar conjugated structure. For example, GO bonding with TiO₂ or WO₃ nanoparticles need to be calcinated at high temperature; thus, GO can form reduced GO (rGO) or other forms of GO. Raj Pant et al. synthesized Ag-ZnO/rGO nanocomposite in an autoclave at 140°C [15], and Gao et al. synthesized sulfonated graphene oxide-ZnO-Ag photocatalyst at 300°C [13]. Only limited reports are available for different fabrication methods of Ag-ZnO/GO to treat organic pollutants under UV light [16].

Herein, we report a novel fabrication method of Ag-ZnO/GO nanocomposite and its characterizations, and then apply it for use in effective removal of the cationic dye methylene blue (MB) in an aqueous solution. This study also aims to save energy and costs by using a novel synthesis method at low temperature, with visible lamps rather than UV lamps as the light source of the photocatalytic process. A suitable photocatalytic degradation mechanism for enhanced MB removal was also proposed.

2. Materials and Methods

2.1. Preparation of Photocatalysts. Sodium hydroxide (NaOH), silver nitrate (AgNO₃), zinc sulfate heptahydrate (ZnSO₄·7H₂O), ascorbic acid (C₆H₆O₆), and MB (C₁₆H₁₈N₃SCl) were purchased from Sigma-Aldrich Co., Ltd. All purchased chemicals were used without any further purification. Distilled water was used for experiments.

ZnO nanoparticles were prepared by dropwise addition of 25 mL of NaOH 0.4 mol/L into 25 mL of ZnSO₄ 0.2 mol/L at an approximate addition rate of 5 mL/min. After stirring with a magnetic stirrer (GLHPS-G, Global Lab., Ltd.) at a speed of 150 rpm for 60 min, the solution was kept at 60°C for 2 h. The Ag-ZnO composite nanoparticles were prepared by adding 6 mL of ascorbic acid 0.01 mol/L and 13 mL of AgNO₃ 0.01 mol/L into the solution of NaOH and ZnSO₄, while stirring under the same condition as in the first experiment, and again the solution was kept at 70°C for 2 h.

For the preparation of Ag-ZnO/GO, the steps of the Ag-ZnO synthesis procedure were repeated. At the same time, 50 mg of GO was mixed into 50 mL water in an ultrasonicator (D250H, DAIHAN Scientific, Co., Ltd.) for 30 min at room temperature. The two solutions were then mixed, and the final solution obtained was kept at 70°C for 2 h. In the last step, the synthesized products were centrifuged and washed with deionized water several times and dried in a vacuum at 70°C for 24 h.

2.2. Characterization. The synthesized reaction products were characterized by X-ray diffraction (XRD; Bruker AXS D8 ADVANCE) to identify the structure and phase composition. Wide-angle patterns were recorded from 2θ = 10° to 80° using a step size of 0.1°. Their surface morphologies and microstructures were examined using field emission scanning electron microscopy (FE-SEM; JEOL, JSM-6500F, 10 kV) and transmission electron microscopy (TEM; JEOL, JEM-2100F). Composition mapping of the major elements on the material surface was carried out using the Brunauer–Emmett–Teller ( BET) method using nitrogen gas adsorption. Room temperature photoluminescence (PL) spectra were recorded using a fluorescence spectrometer.

The visible light absorption of the synthesized products was measured in the range of 400 to 800 nm using a UV-Vis spectrophotometer (GENESYS™ 10S UV-Vis, USA) with integrating sphere accessories. To plot the calibration curve of MB dye, aqueous dye solutions were prepared at a concentration ranging from 1 to 25 mg/L by using distilled water. The concentrations of the MB solutions were determined using the obtained calibration curve. The dye removal efficiency was calculated based on following equation:

\[
\text{Removal efficiency (\%)} = \left(\frac{C_0 - C}{C_0}\right) \times 100\% ,
\]

where \(C_0\) (mg/L) is the concentration of the MB solution at the initial time \(t = 0\) (min) and \(C\) (mg/L) is the concentration after the treatment reaction in a dark condition or after UV-visible light irradiation.

2.3. Photocatalytic Activity Measurement. The photocatalytic activity of the synthesized materials was estimated by using an illumination system consisting of five lamps [visible lamp (EFTR 20EX-D, Kumho Co., Ltd., Republic of Korea)/UV lamp (AL-2220D1 20W, Alim Co., Ltd, the Republic of Korea)] as irradiation sources. In a typical process, 20 mg of the synthesized photocatalyst was suspended in 20 mL of MB
with a concentration of 15 mg/L in a cylindrical glass reactor. Before starting the photocatalytic reaction, the suspension was stirred for 30 min in a dark condition to obtain an adsorption/desorption equilibrium between the dye and the photocatalyst. Photocatalytic reactions were carried out under a stable condition (stirring speed of 80 rpm at intervals of 3 h under light irradiation).

To clarify the dominant radical or ion on the photocatalysis reaction, the experiments using different radical scavengers were performed. Scavengers including tert-butyl alcohol, benzoquinone, ammonium oxalate, and K$_2$S$_2$O$_8$ were used for OH$^\bullet$, O$_2$$^\bullet_2$, holes and electrons, respectively. The experiment was carried out similar to the removal experiment with the added radical scavenger (0.1 mmol).

3. Results and Discussions

3.1. Characterization of Material

3.1.1. XRD Pattern Analysis. Figure 1 showed the XRD patterns of the ZnO and Ag-ZnO/GO samples. The Ag-ZnO/GO spectrum included a diffraction peak at $2\theta = 11.5^\circ$ of pristine GO [17, 18], indicating that the major form of graphene in the synthesized material was GO. In both ZnO and Ag-ZnO/GO spectra, the observed diffraction peaks at $2\theta = 32, 34, \text{ and } 36^\circ$ confirmed the presence of the hexagonal wurtzite structure of ZnO. In general, the intensity of the diffraction peaks decreases greatly with the increase in doping concentration [19]. Thus, it can be observed that the peak intensity of ZnO in the Ag-ZnO/GO sample was decreased as compared with that of the ZnO sample. Besides, the XRD pattern of Ag-ZnO/GO mainly showed a small silver peak at $2\theta = 38^\circ$ [20–22], and this confirmed the doping activity of Ag onto the structure of ZnO. Simultaneously, a small amount of Ag entered the ZnO crystal structure, as confirmed by the broadening at $2\theta = 32, 34, \text{ and } 36^\circ$ in the Ag-ZnO/GO peaks compared to ZnO peaks [23, 24]. The value of full width at half maximum (FWHM) of diffraction peak is showed on Table 1.

![Figure 1: XRD patterns of the ZnO and Ag-ZnO/GO samples.](image)

**Table 1: FWHM in the samples.**

<table>
<thead>
<tr>
<th>Sample</th>
<th>FWHM (cm$^{-1}$)</th>
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<tr>
<td>ZnO</td>
<td>0.2210</td>
</tr>
<tr>
<td>Ag-ZnO/GO</td>
<td>0.2377</td>
</tr>
</tbody>
</table>

The chemical compositions of ZnO and Ag-ZnO/GO were analyzed by energy dispersive spectrometry (EDS) and mapping technique in conjunction with SEM (Figure 3). All the peaks were ascribed to Zn, Ag, and O in the ZnO sample, and peak of C elements appeared in the composite sample. The mapping results confirmed the presence and uniform distribution of zinc, silver, and oxygen on the GO surface. In combination, these elemental mapping and SEM and TEM results demonstrate the capability of GO to function as an effective scaffold for ZnO and silver.

3.1.2. Morphology and Microstructure by SEM and TEM Analysis. FE-SEM and TEM analyses were used to identify the morphology and microstructure of the synthesized materials. The SEM images of ZnO before addition of Ag and GO exhibited high degree of uniformity in the nanosized ZnO particles (Figures 2(c) and 2(d)). In the Ag-ZnO/GO heterostructure, GO sheets were consistently decorated with Ag and ZnO nanoparticles. TEM images revealed a hexagonal shape for ZnO (Figure 2(e)) with a diameter 50–60 nm, which is in good agreement with the diameter of ZnO revealed in SEM images. Figure 2(f) shows the TEM image of Ag-ZnO/GO composite, demonstrating the attendance of few-layered GO sheets decorated with Ag and ZnO particles which may result in better adsorption capacity and electron-hole separation.

3.1.3. XPS Analysis. The surface element composition and chemical state of the as-synthesized samples were analyzed by XPS analysis, as shown in Figure 4. The peaks of Zn 2P$_{3/2}$ and Zn 2P$_{1/2}$ of the synthesized ZnO and Ag-ZnO/GO were observed at around 1,021.8 eV and 1,045.1 eV, respectively, which are very similar to the peaks of pure ZnO (Figure 4(a)). This finding, therefore, demonstrated the presence of the Zn$^{2+}$ form in both samples [19, 25]. In the C1s spectra of Ag-ZnO/GO (Figure 4(b)), the presence of C was attributed to the GO addition. Compared to ZnO, the C peaks of the Ag-ZnO/GO nanocomposite were shifted toward a slightly lower binding energy. In the ZnO sample, the presence of C originated from the vacuum oil used in the pretreatment system before the XPS testing. The four peaks in the Ag-ZnO/GO sample at 282.4, 284.8, 286.1, and 288.6 eV were ascribed to Zn-C, C-C/C$\equiv$C, Zn-O-C, and C=O, respectively [26, 27]. The presence of abundant carbon species on the surface of the Ag-ZnO/GO composite increased the photodegradation because it facilitated the
contact with organic pollutant molecules. The high resolution spectra of the strong O1s peak (Figure 4(c)) at 531.3 eV in the ZnO sample was due to the oxygen in the ZnO crystal lattice (Zn-O bonds) [28, 29]. Two O1s peaks at 531.5 and 531.8 eV in the Ag-ZnO/GO composite sample revealed the presence of surface oxygen complexes in the carbon phase [21, 28]. These oxygen-containing groups increased the photocatalytic activities due to their involvement in the production of active radicals, which play an important role in the photodegradation process. The Ag 3d XPS peaks of Ag-ZnO/GO, shown in Figure 4(d) located at 367.7 and 374.0 eV, were ascribed to Ag 3d5/2 and Ag 3d3/2, respectively [30, 31]. The XPS results were in good agreement with the aforementioned XRD and EDS results. These observations in Figure 4 further confirmed the successful preparation of Ag-ZnO/GO nanoparticles and viability of Ag-ZnO/GO as a superior nanocomposite material.

3.1.4. UV-VIS Reflectance Spectra and Band Gap. The optical absorption properties of the synthesized nanomaterials were investigated by UV-VIS reflectance spectra (Figure 5). The doping activity induced a shift from the UV light absorption of ZnO to the visible light absorption of Ag-ZnO/GO. The optical band gaps of the synthesized materials were calculated by using the following Tauc equation:

$$ ahv = A(hv - E_g)^n, $$

where $\alpha$ is the absorption coefficient, $E_g$ is the band gap, $A$ is a constant, and $n$ is an index that characterizes the optical absorption process (for direct band gap semiconductor material $n = 1/2$). By extrapolating the linear region of the plot $(ahv)^2$ vs. $hv$, the band gap could be estimated. The band gap values for ZnO and Ag-ZnO/GO are given in Figure 5. Band gap of Ag-ZnO/GO (2.92 eV) is smaller than that of synthesized ZnO (3.15 eV). This decreased band gap may have been due to the introduction of silver and carbon as dopants in the ZnO lattice. Similar phenomena have been observed in ZnO-based material systems in other studies [19, 32, 33].

3.1.5. BET Surface Area. The specific surface area of the synthesized materials was measured using the BET method with N2 adsorption-desorption (Figure 6). The identified surface area of Ag-ZnO/GO was almost 3.6 times larger than

Figure 2: FE-SEM and TEM analyses of synthesized materials.
that of ZnO. With Ag-ZnO/GO present in a dark condition, pollutant adsorption is mainly assisted by the increased specific surface area ($S_{\text{BET}}$). In general, graphene has a very high specific surface area [34], and thus, it could provide a high adsorption capacity. GO, the oxidized form of graphene, contains oxygen functional groups on its surface that can become adsorption sites. Therefore, the enhanced degradation capacity under visible light can be attributed to the adsorption power of GO combined as a semiconductor or adsorption substrate [35, 36]. In addition, the increased pore size of Ag-ZnO/GO nanocomposite could lead to the increases in adsorption efficiency.

3.1.6. PL Spectra. The PL spectra of as-prepared ZnO and Ag-ZnO/GO at room temperature are presented in Figure 7. As observed, the PL intensities of the samples increase in the following order: Ag-ZnO/GO and ZnO. The PL intensity of the composite sample is weaker as compared with that of the ZnO sample, indicating that the fluorescence of the composite is quenched more efficiently than that of ZnO. It also indicated that the incorporation of ZnO with Ag and GO can improve the separation of photoinduced electrons and holes. Thus, there is a high agreement with the order of PL intensities when compared with the result from the removal experiments, and the recombination process can be
Figure 4: XPS analyses of synthesized materials.
significantly suppressed through the combination of ZnO with Ag and GO.

3.2. Removal of MB Dye Using Synthesized Materials. It is commonly accepted that most dyes are resistant to biodegradation and direct photolysis, and many N-containing dyes such as MB undergo natural reductive anaerobic degradation to yield potentially carcinogenic aromatic amines [31]. In this study, therefore, MB was chosen as a model contaminant to evaluate the photocatalytic activity of the synthesized photocatalysts.

Figures 8(a) and 8(b) show the UV-Vis absorption spectrum and removal efficiency of MB degraded by using synthesized materials under dark and light (visible and UV) irradiation conditions, respectively. ZnO did not show any significant adsorption of MB, when the addition of Ag-ZnO/GO into the MB solution without any light source afforded an MB removal efficiency of around 20%. After the adsorption, visible light or UV light was directed at the MB removal system containing Ag-ZnO/GO added into MB solution as a photocatalyst. The addition of Ag-ZnO/GO into the MB solution under visible light and UV light irradiation increased the MB removal efficiency after 3h by up to 85% and 99%, respectively. The comparison of the light absorption results between the dark and light irradiation conditions clearly demonstrated that most of the MB removal effects were due to photocatalytic degradation by the Ag-ZnO/GO nanocomposite. Under visible light, MB removal was significantly increased by Ag-ZnO/GO because of the combination effects of the adsorption and photocatalytic degradation. Under UV conditions, removal efficiency reached up to 99% because of the high photon energy in UV light, and so photodegradation could occur more strongly than under the visible light. The photocatalytic activity by Ag-ZnO/GO under visible light is explained in the mechanism section.

To clarify the role of photogenerated radical species in the removal process, different scavengers were used. It is observed from the scavenger tests’ result that the degradation level of MB was significantly inhibited when tert-butyl
alcohol and ammonium oxalate were added. Thus, it is clear that most of the reactive radicals responsible for catalytic activity are found to be OH• and photogenerated holes (Figure 9).

Synthesized composite material have higher surface area and greater numbers of active sites as compared with ZnO, where the photogenerated charge carriers react with absorbed molecules to form hydroxyl and superoxide radicals. A set of experiments was carried out in order to check the reusability and stability of the composite catalysts. The photocatalytic activities were almost stable in the first 4 cycles. From the 5th cycle, the removal of MB was decreased; it might be due to the loss of adsorption properties after several centrifugation and cleaning process. The particles readily form aggregates, leading to the loss of the original structure and active sites, thus decreasing the photocatalytic efficiency.

3.3. Effect of Initial Process Parameter. pH of the solution has been reported as one of the most important factors affecting the removal efficiency of organic pollutants by photocatalytic processes in an aqueous solution [37, 38]. Interpreting the pH effects on the MB dye removal process is a difficult task because it is affected by multiple factors. The effect of pH on the removal of MB dye was investigated in the pH range 3 to 12. The pH of the point of zero charge (pH pzc) of ZnO was about 8.6 [39]. At pH above pH pzc, the surface of the ZnO particles was mostly positively charged. As the solution pH increases from the acidic range up to pH pzc of ZnO (pH < 8.6), the decreased H3O+ concentration produces less repulsion of Ag-ZnO/GO with the positively charged MB molecules, resulting in increased adsorption of MB. As the solution pH further increases above pzc (pH > 8.6), the increased OH− produces more electron repulsion of Ag-ZnO/GO with negatively charged MB molecules, leading to less adsorption. Therefore, pH 8.5–9 was chosen as the optimal pH for MB adsorption (Figure 11(a)).

Figure 11(b) shows the effects of different Ag-ZnO/GO loadings on the MB removal process under visible light irradiation. As the dosage of Ag-ZnO/GO increased up to 1.0 g/L, the MB removal effect also increased. The increased Ag-ZnO/GO dosage led to more active sites for adsorption and thus more moiety availability for photocatalytic degradation of MB molecules. However, even the MB removal efficiency decreased as the dosage loading was increased above 1 g/L. At higher dosages, there was excessive increase in the amount of suspended Ag-ZnO/GO, with excessive addition, disturbing the penetration of visible light into the reaction system. This also led to reduction in the generation
of the electron-hole pairs and subsequent reduction in the production of oxy-radicals and hydroxyl radicals [40]. Furthermore, excessive photocatalyst dosage increases the pollutant removal costs. Hence, 1 g/L was determined to be the optimum Ag-ZnO/GO dosage.

Different initial MB solution concentrations, ranging from 1 mg/L to 25 mg/L, were used to evaluate the MB removal effect by Ag-ZnO/GO (Figure 11(c)). The MB removal efficiency decreased when the initial MB concentration was more than 15 mg/L within 3 h of irradiation. When the MB concentration was beyond the limit of 15 mg/L, the MB molecules adsorbed on the adsorbent/photocatalyst surface repulsed further MB molecules from approaching the adsorbent/photocatalyst, thereby decreasing MB removal. In addition, a high initial MB concentration hindered visible light penetration due to increased turbidity, as explained in the previous section, which decreased the light irradiation effect for photocatalytic degradation of MB [41, 42].

### 3.4. Photocatalytic Mechanism

Three mechanisms proposed to explain the increased photocatalytic degradation of MB dye by Ag-ZnO/GO under visible light irradiation are schematically shown in Figure 12.

The first proposed mechanism for the increased MB removal is associated with GO addition to the photocatalytic system (Figure 12(a)). GO was used as a better substrate for the photocatalytic reaction by increasing the surface area of the photocatalyst. Moreover, the photocatalytic degradation efficiency of MB by Ag-ZnO/GO was improved by combining it with a zero band gap semiconductor, GO [43, 44]. Some previous studies have reported that GO can also enhance the photocatalytic ability of ZnO under visible light irradiation due to resonance effects, including the increased surface area with added GO and increased formation of π-π* interactions between the dye molecules [25]. The high surface area of GO can contribute to the effective adsorption of MB molecules on the photocatalyst surface. MB is a sensitive chromophore that absorbs light in a wide range of wavelengths, including the visible region [45, 46], and thus, MB molecules easily enter an excited status. The electrons in the excited MB* can jump to the conduction band (CB) of ZnO through GO [47] and then be transferred to various Ag levels (Figures 12(b) and 12(c)). This series of excited electron transfer can minimize or delay the recombination of electrons with holes. Therefore, the excited electrons can have more delayed recombination, while simultaneously increasing the charge transfer capacity from the valence band (VB) to the CB of ZnO.

The second mechanism for the enhanced photocatalytic degradation of MB could be due to the Ag doping effect into the ZnO crystal lattice (Figure 12(c)). It is well known that band gap is a region of energy with no allowed states. The density of states versus energy depends on the chemical composition of the material, and the state density distribution will be changed if the chemical composition is changed. In this case, Ag dopant is the impurity, so the chemical composition was changed by doping. When the doping density is high enough, the dopant states generate a band. If this band is very close to the valence or conduction band edge, the band gap will decrease. The electrons transferred to the CB of ZnO tend to be transferred to Ag at that time, which prevents delay of the recombination of the excited electrons and holes. Addition of Ag led to the formation of "stairs" that allow the excited electrons to move easily to higher energy levels with visible light irradiation rather than directly moving down to the holes. The minimized recombination of the excited electrons in the CB with the holes in the VB can increase the opportunity for the production of oxy-radicals by reaction with O₂ molecules, leading to the oxidative degradation of MB molecules.

The third proposed mechanism is based on the narrowed band gap of the semiconductor (Figure 12(b)). The major limitation of ZnO is its restriction to UV light irradiation because of its wide band gap. This weakness was improved through Ag doping into the ZnO lattice by narrowing the band gap. Dotted green lines (Figure 12(c)) represent a new band gap for ZnO, which was narrowed by the interaction between ZnO, Ag, and GO during the synthesis of the Ag-ZnO/GO nanocomposite [48]. The major oxidative and reductive processes for the photodegradation of MB by using Ag-ZnO/GO with a narrowed band gap under visible light illumination can be explained as shown in equation (3) to (11).

$$\text{Light with appropriate spectrum } + \text{Ag-ZnO/GO} \rightarrow \text{Ag-ZnO/GO} (h^+ + e^-)$$

(3)
Figure 11: Effect of the initial parameter on the MB removal efficiency.

Figure 12: Proposed mechanism for MB removal using synthesized material.
Table 2: Comparison between previous and current studies for MB removal using photocatalytic degradation.

<table>
<thead>
<tr>
<th>Photocatalyst</th>
<th>Chemical ingredients</th>
<th>Calcination temperature</th>
<th>Light source</th>
<th>Adsorption</th>
<th>Photocatalysis</th>
<th>Total removal</th>
<th>Remark</th>
<th>Ref.</th>
</tr>
</thead>
<tbody>
<tr>
<td>Sulphonated GO-ZnO-Ag</td>
<td>Zn(CH$_3$COO)$_2$.2H$_2$O, HMTA, EG, AgNO$_3$, HCl, ClICH$_2$CH$_2$SO$_2$H, AgNO$_3$, NaNO$_3$, KMnO$_4$, H$_2$O$_2$, H$_2$SO$_4$</td>
<td>160</td>
<td>Visible</td>
<td>20%</td>
<td>78%</td>
<td>98%</td>
<td>Sulphonated GO high calcination temperature</td>
<td>[13]</td>
</tr>
<tr>
<td>Ag-ZnO/ RGO</td>
<td>GO, bis-hexamethylene triamine, Zn(NO$_3$)$_2$.6H$_2$O, AgNO$_3$, ethanol</td>
<td>140, in autoclave</td>
<td>Visible</td>
<td>Not separated</td>
<td>65%</td>
<td>Reduced GO high calcination temperature in autoclave</td>
<td>[15]</td>
<td></td>
</tr>
<tr>
<td>Graphene-Ag/ZnO</td>
<td>Graphene, EG, CH$_3$COOAg, Zn(CH$_3$COO)$_2$.2H$_2$O, NaOH</td>
<td>160</td>
<td>Visible</td>
<td>28.6%</td>
<td>65.6%</td>
<td>94.4%</td>
<td>Graphene expensive high calcination temperature in autoclave</td>
<td>[8]</td>
</tr>
<tr>
<td>Ag/ZnO/GO</td>
<td>Graphite oxide, ZnO, AgNO$_3$</td>
<td>55</td>
<td>UV</td>
<td>Not separated</td>
<td>98%</td>
<td>GO UV light high energy temperature</td>
<td>[16]</td>
<td></td>
</tr>
<tr>
<td>Ag-ZnO/GO</td>
<td>Graphene oxide, AgNO$_3$, ZnSO$_4$.7H$_2$O, C$_6$H$_6$O$_6$</td>
<td>70</td>
<td>Visible/UV</td>
<td>25%</td>
<td>60%</td>
<td>85%</td>
<td>GO visible light low energy high safety issue</td>
<td>This study</td>
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<tr>
<td></td>
<td>Graphene oxide, AgNO$_3$, ZnSO$_4$.7H$_2$O, C$_6$H$_6$O$_6$</td>
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<td>Visible/UV</td>
<td>25%</td>
<td>74%</td>
<td>99%</td>
<td>GO visible light low energy high safety issue</td>
<td>This study</td>
</tr>
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(I) Oxidative reactions with holes
\[ h^+ + H_2O \rightarrow H^+ + OH^* \] (4)
\[ 2h^+ + 2H_2O \rightarrow 2H^+ + H_2O_2 \] (5)
\[ H_2O_2 \rightarrow HO^* + \cdot OH \] (6)

(II) Reduction reaction with O$_2$
\[ 2e^- + O_2 \rightarrow \cdot O_2^- \] (7)
\[ \cdot O_2^- + 2H^+ \rightarrow H_2O_2 + O_2 \] (8)
\[ H_2O_2 \rightarrow HO^* + \cdot OH \] (9)

(III) Photocatalytic oxidation with oxy-radicals
\[ \text{Dye/Dye}^+ + \cdot OH \rightarrow \text{final products (CO}_2, H_2O, \ldots) \] (10)
\[ \text{Dye/Dye}^+ + \cdot O_2^- \rightarrow \text{final products (CO}_2, H_2O, \ldots) \] (11)

3.5. Energy and Cost Issue. Nowadays, the demand and market for the use of nanoparticles or nanocatalysts in pollutant removal are increasing. As discussed above, ZnO nanoparticle is one of the most promising materials for wastewater treatment. Performance of ZnO can be enhanced by adding some ingredients to make better nanocomposite. The methods currently developed for making better ZnO nanomaterials mainly consist of sol-gel template and hydrothermal methods. However, the requirement of high crystallinity is a major problem in ZnO synthesis. With the sol-gel method, calcination of gels or thermal annealing of emulsions is therefore required to induce crystallization of the nanoparticle, and thus, normally a high temperature of more than 200°C is required. Hydrothermal methods are directly carried out at slightly lower temperatures than sol-gel methods (but not less than 120°C). However, nanocrystals formed with hydrothermal methods agglomerate and thus are insoluble in most solvents, and thus, some stabilizing agents are required to prevent agglomeration. Their characteristics from previous relevant study outcomes are summarized in Table 2 with comparisons.

The search for a simple and economic synthesis method to derive nanoparticles with good size and shape at low temperature is still an open challenge. The ability to produce nanomaterials at lower temperatures is needed for the purpose of saving energy and increasing safety for large-scale production. In this study, we demonstrate that it is possible to cost effectively produce nanomaterials at low temperature in considerable quantities, with increased safety in a wide range of applications. By adding the noble metal Ag as a dopant and GO as a high surface area adsorption substrate, firstly, our nanomaterial can work excellently for adsorption and also as a photocatalyst, even under visible light. Researchers have previously used UV light in their studies to irradiate the photocatalyst, and the removal efficiency of their process was very high. However, the price of a UV lamp is at least two times as high as the price of a visible lamp. Furthermore, UV waves are invisible but very harmful for human eyes. Secondly, for processing our current method, only a calcination temperature of around 70°C is needed without any requirements for complex instruments. From an economical view, such fabrication may offer better opportunities to significantly lower the cost of manufacturing nanomaterials, while bringing environmental advantages such as low energy consumption and reduced CO$_2$ emissions. Thirdly, the simplicity of the synthesis procedure...
would make it safe for workers and easy to apply to industrial manufacturing.

4. Conclusion

Ag-ZnO/GO nanocomposite was successfully synthesized by facile aqueous solution reactions at low temperature. The MB removal efficiency increased up to 99% under the UV light and 85% under visible light. The optimum conditions for maximum removal efficiency of MB were pH 8.5–9, temperature 35°C and dosage 1 g/L at MB concentration 15 mg/L. The significant increase in photocatalytic degradation for MB removal exhibited by Ag-ZnO/GO was due to the combined effects of the two semiconductors, ZnO and GO, and Ag doping into the ZnO crystal lattice. The proposed mechanism for enhanced removal includes an increase in adsorption by adding GO with a high surface area and an increase in photocatalytic activities due to improved charge transfer achieved through lowering the band gap energy of ZnO, thus minimizing the recombination of the excited electrons in the CB with the holes in VB of ZnO, leading to higher removal rate of MB.

Data Availability

The data used to support the findings of this study are available from the corresponding author upon request.

Conflicts of Interest

The authors declare that they have no conflicts of interest.

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