

Research Article

Rapid Determination of Mercury Ions in Environmental Water Based on an N-Rich Covalent Organic Framework Potential Sensor

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In this article, an N-rich covalent organic framework (COFTFPB-TZT) was successfully synthesized using 4,4',4'-(1,3,5-triazine-2,4,6-triyl) trianiline (TZT), and 4-[3,5-bis(4-formyl-phenyl) phenyl] benzaldehyde (TFPB). The as-prepared COFTFPB-TZT possesses irregular cotton wool patches with a large specific surface area. A novel selective electrode based on COF_{TFPB-TZT} was used for the determination of Mercury ions. The abundance of N atoms in COF_{TFPB-TZT} provides more coordination sites for Hg²⁺ adsorption, resulting in a change in the surface membrane potential of the electrode to selectively recognize Hg²⁺. Under optimal experimental conditions, the ion-selective electrode shows a good potential response to Hg²⁺, with a linear range of $1.0 \times 10^{-9} \sim 1.0 \times 10^{-4}$, a Nernst response slope of 30.32 ± 0.2 mV/-PC at 25°C and a detection limit of 4.5 pM. At the same time, the mercury-ion electrode shows a fast response time of 10 s and good reproducibility and stability. The selectivity coefficients for Fe²⁺, Zn²⁺, As³⁺, Cr⁶⁺, Cu²⁺, Cr³⁺, Al³⁺, Pb²⁺, NH₄⁺, Ag⁺, Ba²⁺, Mg²⁺, Na⁺, and K⁺ are found to be small, indicating no interference in the detection system. The proposed method can be successfully applied to the determination of Hg²⁺ in 3 typical environmental water samples, with a recovery rate of 98.6–101.8%. In comparison with the spectrophotometric method utilizing dithizone, the proposed method is simple and fast and holds great potential application prospects in environmental water quality monitoring and other fields.

1. Introduction

Heavy metal pollution poses a great threat to human health and causes serious environmental problems. Mercury (Hg) is a highly toxic heavy metal element widely distributed in air, water, and soil [1]. Trace amounts of Hg²⁺ can produce strong toxicity and accumulate in the environment and organisms and can be transferred to the human body through the food chain, directly damaging the human brain, nervous system, kidney, and endocrine system [2, 3]. Therefore, the determination of Mercury ions is necessary in environmental monitoring and clinical analysis, and it is of great significance to improve the quality of human life and protect the Earth's environment. At present, mature methods for mercury-ion detection are as follows: spectrophotometry [4], inductively coupled plasma-mass spectrometry (ICP-MS) [5], atomic absorption/emission spectrometry (AAS/AES) [6], atomic fluorescence spectrometry (AFS) [7], high-performance

liquid chromatography (HPLC) [8], inductively coupled plasma atomic emission spectrometry (ICP-AES) [9], and fluorescence probe detection [10]. Although these methods show excellent detection performance, most of them require expensive precision instruments, complex sample preparation processes, and skilled operators and cannot be conveniently used online or outdoors. Therefore, it is very important to find a simple, rapid, and inexpensive method for the determination of mercury-ion content.

The ion-selective electrode (ISE) method, as a rapid analytical tool, has been widely developed in the fields of the environment, medicine, and agriculture due to its advantages of fast speed, simple preparation, low cost, and high sensitivity [11]. In the process of preparing ion-carrier membranes with different selective electrodes, many organic, and inorganic complexes can be used as selective carriers of Mercury ions, such as crown ether [12], amide [13], glycidyl diamine [14], metalloporphyrin [15], Schiff

base [16], calixarene derivatives [17, 18], and alkylamine Mercury salt complexes [19]. Therefore, there is an urgent need to prepare Mercury ion-selective electrodes with good selectivity and a low detection limit. A large number of studies have proven that compounds containing N, O, and S atoms can form coordination compounds with heavy metal ions [20]. Mercury (II) shows a strong affinity with ligands containing sulphur atoms, such as sulfhydryl compounds [21], sulphur-containing heterocyclic rings [22], and thio-urea derivatives [23], and its interaction relationship has been further verified by density functional theory calculations [24]. Many compounds containing sulphur atoms have been successfully used as active carriers for Mercury ion-selective electrodes [25]. Li et al. [26] prepared a simple and rapid Mercury ion selective electrode based on a 1-undecanethiol (1-UDT) PVC film electrode for the selective determination of Hg^{2+} ions, with a Nernst-response range and a detection limit of 4.5 nM. Fang et al. [27] reported a portable Hg^{2+} nanosensor using the CuS nanozyme functions as a Hg^{2+} recognition unit, which exhibits high sensitivity with a minimum detectable Hg^{2+} concentration of 50 ppt. Diamantis [28] et al. reported a microporous 8-connected Zr^{4+} metal-organic framework (MOF) based on a terephthalate ligand decorated with a chelating 2-picolylamine side group (dMOR-2), which shows a highly efficient fluorescence sensing and sorption of Mercury ions (Hg^{2+}), and the limits of detection were determined to be below 2 ppb for Hg^{2+} .

The covalent organic framework (COF) is a porous conjugated polymer that shows a large surface area, adjustable pore order, and easy functionalization [29–31]. In this study, a COF with an irregular cotton wool patch structure was synthesized from 4,4',4'-(1,3,5-triazine-2,4,6-triyl) trianiline (TZT), and 4-[3,5-bis (4-formyl-phenyl) phenyl] benzaldehyde (TFPB). The as-prepared COF has ordered holes and a large specific surface area, and more N atoms will be exposed on the inner wall of the holes and the specific surface area of COF, providing more adsorption sites for the binding of Hg^{2+} . SEM and TEM characterization also prove the existence of such a recognition effect. The experimental results show that the electrode has high sensitivity, selectivity, reproducibility, and stability. The electrode is used for the rapid determination of Mercury ions in 3 typical environmental water samples and the results are compared with the spectrophotometric method with dithizone. The as-prepared $\text{COF}_{\text{TFPB-TZT}}$ shows excellent selectivity for Hg^{2+} , holding important application value in environmental monitoring, biomedical and agricultural product safety detection, and other fields.

2. Experimental Section

2.1. Reagents and Instruments. 4,4',4'-(1,3,5-triazine-2,4,6-triyl) trianiline (TZT) and 4-[3,5-bis (4-formyl-phenyl) phenyl] benzaldehyde (TFPB) were obtained from Jilin Chinese Academy of Sciences Technology Co., Ltd. (Beijing, China). o-Dichlorobenzene (o-DCB), butyl alcohol (n-BuOH), dichloromethane (DCM), tetrahydrofuran (THF), N, N-dimethylformamide (DMF), acetic acid

(AcOH), and Mercury chloride (Hg_2Cl) were purchased from Aladdin Biochemical Technology Co. Ltd. (Shanghai, China). Potassium ferricyanide ($\text{K}_3[\text{Fe}(\text{CN})_6]$), potassium ferrocyanide ($\text{K}_4[\text{Fe}(\text{CN})_6]$), sodium chloride (NaCl), disodium hydrogen phosphate ($\text{Na}_2\text{HPO}_4 \cdot 7\text{H}_2\text{O}$), sodium dihydrogen phosphate ($\text{NaH}_2\text{PO}_4 \cdot 7\text{H}_2\text{O}$), sodium sulphate (Na_2SO_4), chlorine (HCl), sulfuric acid (H_2SO_4 , 98%), and hydrogen peroxide (H_2O_2 , 30%) were purchased from Sinopharm Chemical Reagents Co. Ltd. (Shanghai, China). All chemicals used in the experiment were of analytical grade. Ultrapure water with a specific resistance larger than $18.3 \text{ M}\Omega \cdot \text{cm}^{-1}$ was used throughout the experiments.

The pH values were measured with a REXPHSJ-4A pH Meter (INESA Instrument Co. Ltd., China). All electrochemical measurements were performed using a CHI-760B workstation (Shanghai Chenhua Instruments Co., China). The potential values were detected with an Ollie Dragon Model 868 pH/mV meter (Thermo Orion, USA). Scanning electron microscopy (SEM) images were obtained using a Nova Nano230 (Thermo Fisher Scientific Co. Ltd., USA). Transmission electron microscopy (TEM) images were obtained using an FEI (Tecnaï G2 F20 TMP, USA). The X-ray powder diffraction (XRD) spectrum was measured using a D/max-2500 diffractometer with a Cu $\text{K}\alpha$ radiation source ($\lambda = 1.54056 \text{ \AA}$), Rigaku Co., Ltd., Tokyo), and a gas adsorption instrument (Micromeritics ASAP 2460 Sorptometer, MAC Instruments Co., Ltd., USA) was used to characterize the samples.

2.2. Synthesis of $\text{COF}_{\text{TFPB-TZT}}$ Material. First, 0.1 mol of TFPB and 0.1 mol of TZT were dissolved in 3 mL of dichloromethane (DCM)/n-butanol mixed solvent (1:1), followed by ultrasound treatment for 30 s. Next, 0.2 mL of 6.0 M acetic acid (AcOH) was added and poured into a Teflon-lined autoclave. Next, the reactor was heated at 120°C for 3 days. The precipitate obtained was repeatedly washed with dimethylformamide until the upper liquid became colourless. Meanwhile, the product was further purified by soaking in THF for 6 hours.

2.3. Electrode Preparation of $\text{COF}_{\text{TFPB-TZT}}$. A “piranha” solution was prepared by mixing 98% H_2SO_4 and 30% H_2O_2 at a volume ratio of 3 : 1 before being poured onto the surface of a gold plate electrode (GPE) for 1 min. The GPE was cleaned with ultrapure water and ethanol and dried with nitrogen for later use. Ethanol was ultrasonically dispersed with 1.5 mg $\text{COF}_{\text{TFPB-TZT}}$, and then the gold surface was immersed in 1.5 g/L $\text{COF}_{\text{TFPB-TZT}}$ ethanol solution and self-assembled at 4°C for 24 h. Finally, the $\text{COF}_{\text{TFPB-TZT}}$ -modified electrode was washed with ethanol and ultrapure water, dried, and stored for later use.

3. Results and Discussion

3.1. Synthesis and Characterization of $\text{COF}_{\text{TFPB-TZT}}$. The morphology and structure of as-prepared $\text{COF}_{\text{TFPB-TZT}}$ before and after reacting with Mercury ions (Hg^{2+}) in solution was further characterized by transmission electron

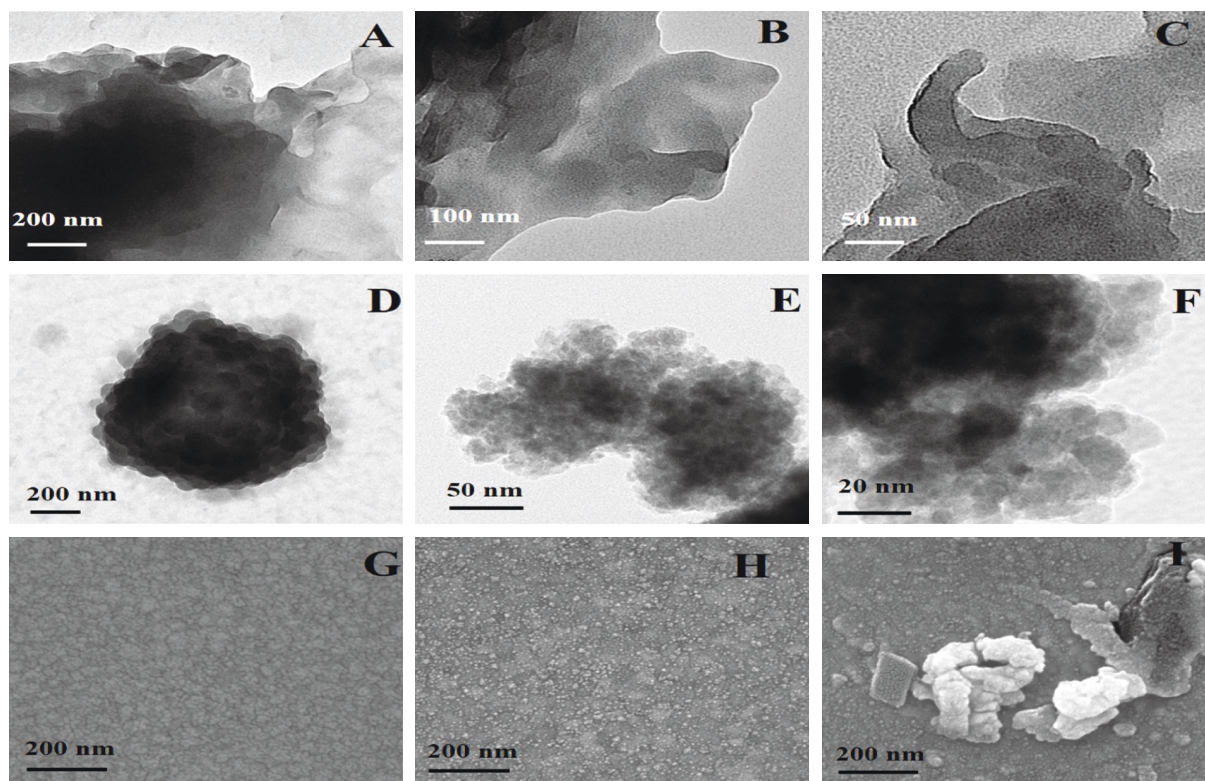
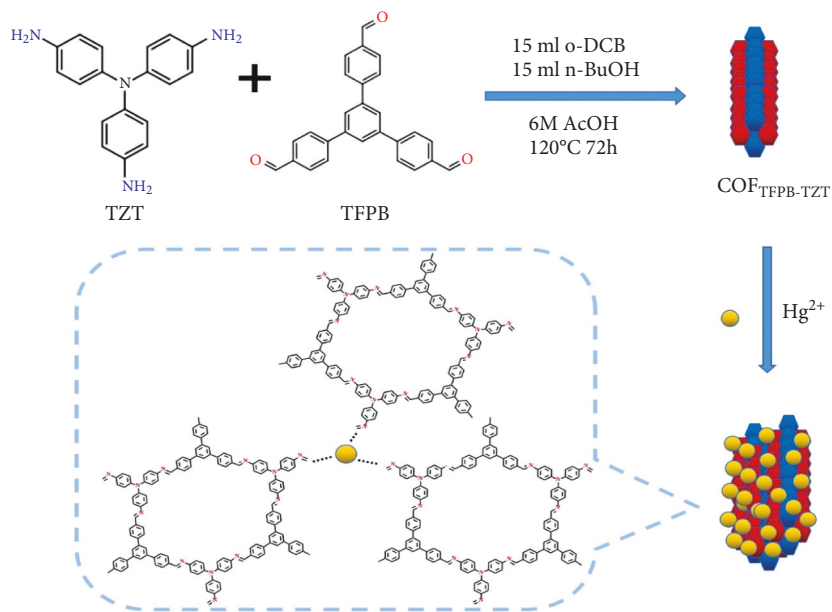


FIGURE 1: TEM images of COF_{TFPB-TZT} (a-c) and COF_{TFPB-TZT}/Hg²⁺ (d-f) at different magnifications. SEM images of bare GPE (g), GPE/COF_{TFPB-TZT} (h), and GPE/COF_{TFPB-TZT}/Hg²⁺ (i).



SCHEME 1: Schematic illustration of the response process of the electrode.

microscopy (TEM). As shown in Figure 1, COF_{TFPB-TZT} exhibits irregular cotton wool patches (Figures 1(a)–1(c)). Upon the incorporation of Hg²⁺, an amount of target Hg²⁺ is adsorbed onto the surface of irregular cotton wool patches via chemical chelation (Figures 1(d)–1(f)), and the irregular cotton wool patches are regathered into stacked layers of nano cotton. The change in the morphology of COF_{TFPB-TZT}

before and after the addition of Hg²⁺ is ascribed to the interaction between COF_{TFPB-TZT} and Hg²⁺ via the chelation of chemical bonds. In addition, the surface morphology characteristics of the GPE/COF_{TFPB-TZT} electrode before and after reacting with Mercury ions were tested by SEM, as shown in Figure 1(g)–1(i). The naked gold plate electrode surface is observed to be very smooth (Figure 1(g)), and after

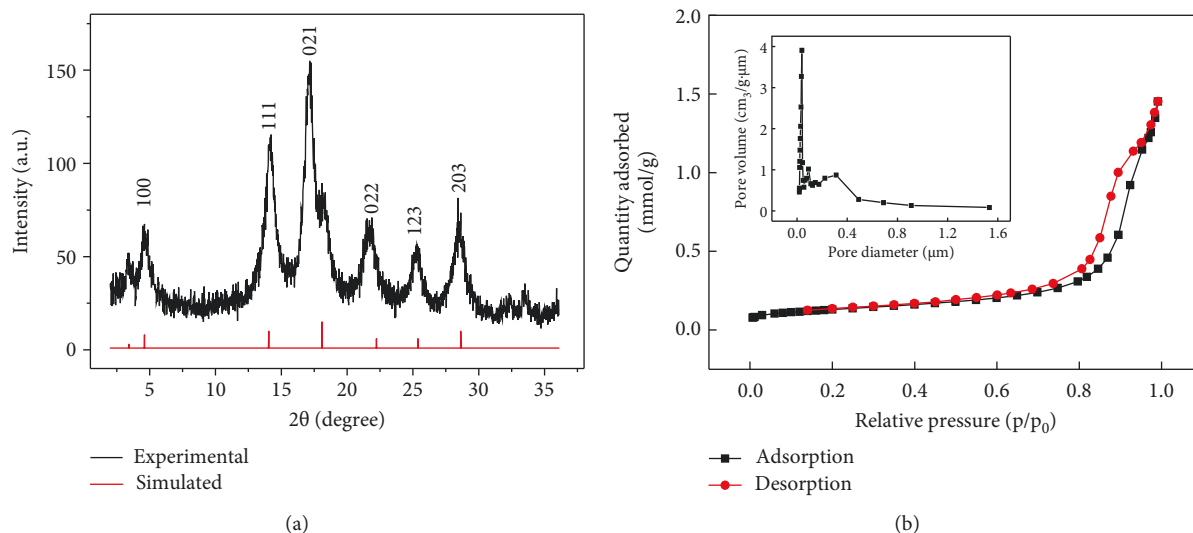


FIGURE 2: (a) Experimental (black curve) and simulated (red curve) XRD patterns for the as-prepared $\text{COF}_{\text{TFPB-TZT}}$. (b) N_2 adsorption-desorption isotherm (BET) for the as-prepared $\text{COF}_{\text{TFPB-TZT}}$. The insert shows the pore distribution of $\text{COF}_{\text{TFPB-TZT}}$.

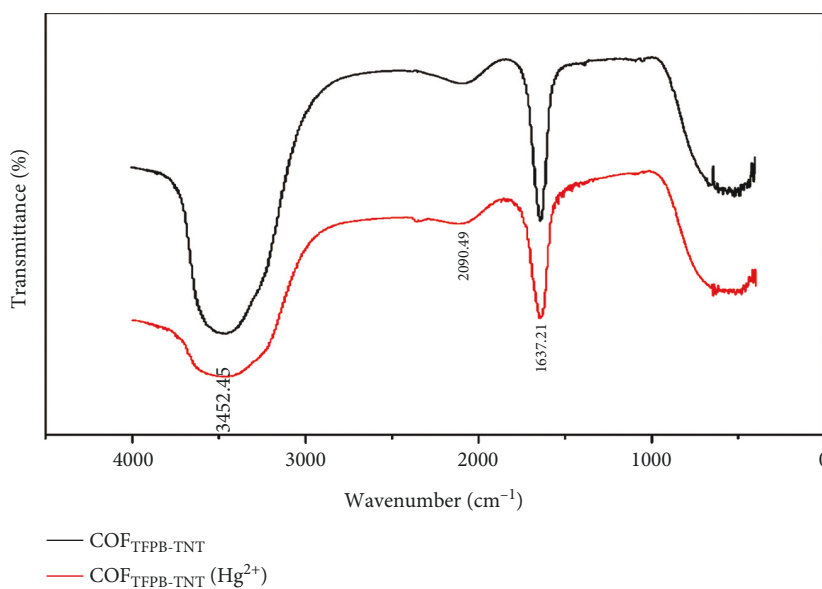


FIGURE 3: FT-IR spectra for the as-prepared $\text{COF}_{\text{TFPB-TZT}}$ (black curve) and $\text{COF}_{\text{TFPB-TZT}}/\text{Hg}^{2+}$ (red curve).

the self-assembly of $\text{COF}_{\text{TFPB-TZT}}$, the surface coverage becomes rough (Figure 1(h)). After testing in mercury-ion solution, the surface of the as-prepared $\text{COF}_{\text{TFPB-TZT}}$ changes significantly, and the electrode appears as larger agglomerates (Figure 1(i)). As shown in Scheme 1, the change in the structure can be ascribed to Mercury ions and amino functional groups of $\text{COF}_{\text{TFPB-TZT}}$ in the formation of strong coordination bonds, resulting in the complexes gathering strongly and changing the shape of the surface. Thus, different electron microscope images are presented.

The crystalline structure of $\text{COF}_{\text{TFPB-TZT}}$ was evidenced by X-ray diffraction (XRD) (Figure 2(a)), and the corresponding characteristic diffraction peaks for $\text{COF}_{\text{TFPB-TZT}}$ for the (100), (111), (021), (022), (023) and (203) planes are observed, which is in agreement with the simulated

spectrum. In addition, adsorption-desorption isotherms (BET) were obtained to evaluate the specific surface area and pore size of the as-prepared $\text{COF}_{\text{TFPB-TZT}}$. The specific surface area of $\text{COF}_{\text{TFPB-TZT}}$ is deduced to be $80.9 \text{ m}^2 \cdot \text{g}^{-1}$, and the pore size of $\text{COF}_{\text{TFPB-TZT}}$ is 0.35 nm (Figure 2(b)), providing more active sites for Hg^{2+} ions. The XRD and BET results confirm that the as-prepared $\text{COF}_{\text{TFPB-TZT}}$ possesses a highly crystalline structure and ordered holes with a large specific surface area, providing more adsorption sites for the binding of Hg^{2+} . The N-atom rich composition of $\text{COF}_{\text{TFPB-TZT}}$ is responsible for the formation of $\text{COF}_{\text{TFPB-TZT}}/\text{Hg}^{2+}$, as evidenced by the FT-IR spectrum. As shown in Figure 3, the characteristic peak at 2390.49 cm^{-1} further proves the coordination relationship between the N element and Hg^{2+} . These results confirm that an N-rich covalent organic

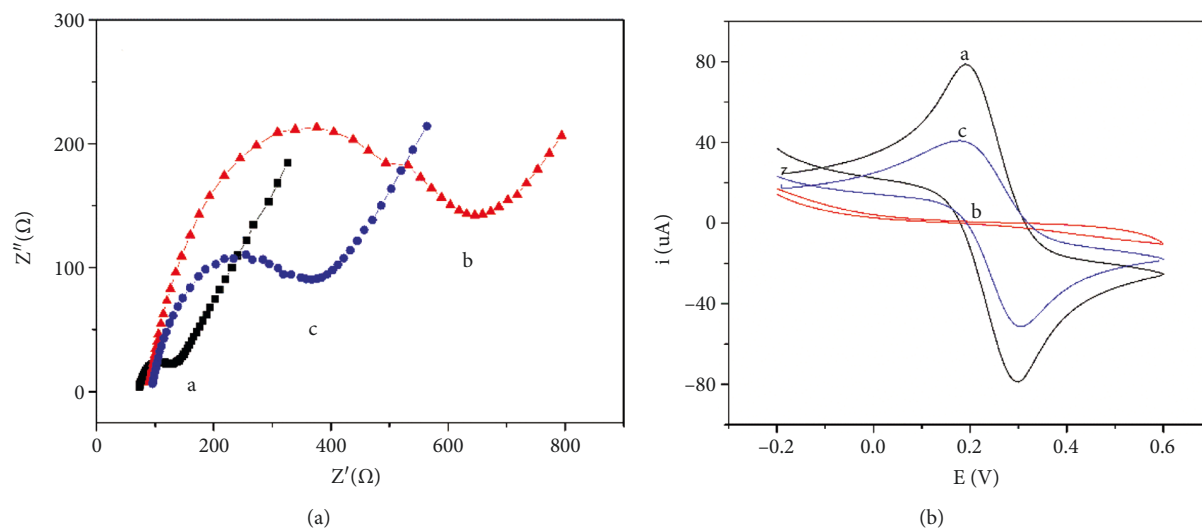


FIGURE 4: Electrochemical characteristic curves measured with impedance spectroscopy (a) and cyclic voltammetry (b) for bare GPE (A), GPE/COFTFPB-TZT (B), and GPE/COFTFPB-TZT/Hg²⁺ (C) in a media solution containing 2.0 mmol/L K₃Fe (CN)₆ and 0.2 mol/L Na₂SO₄.

framework (COF_{TFPB-TZT}) can be successfully synthesized with 4,4',4'-(1,3,5-triazine-2,4,6-triyl) trianiline (TZT), and 4-[3,5-bis (4-formyl-phenyl)phenyl]benzaldehyde (TFPB) reagent through the dehydration condensation reaction of amine aldehydes.

3.2. Electrochemical Characterization of COF_{TFPB-TZT}. To verify the interaction between the carrier and Mercury ion, the electrochemical behaviour of the different modified electrodes above was investigated by electrochemical impedance spectroscopy (EIS) and cyclic voltammetry (CV). As shown in Figure 4(a), curve a represents the bare gold plate electrode, and the impedance value is very small, indicating that the pretreated bare gold plate electrode has a strong ability to transfer electrons. Curve b represents the COF_{TFPB-TZT}-modified gold plate electrode. The semicircle appears in the high frequency part, and the impedance value increases significantly, indicating that COF_{TFPB-TZT} forms a nonconductive monomolecular self-assembled film on the gold surface through the sulfhydryl group at one end, which hinders the electron conduction of [Fe (CN)₆]^{3-/4-} on the electrode surface. After combining Hg²⁺ (1.0 × 10⁻⁴ mol/L), the impedance value of the electrode decreases correspondingly (curve c), which is due to the strong coordination between the N element group and Hg²⁺ at the other end of COF_{TFPB-TZT}, which adsorbs positively charged Mercury ions and enhances the electronic conductivity, resulting in an electrochemical conduction current and reduced interfacial impedance. The variation trend for the impedance can also be verified by the corresponding cyclic voltammetry (Figure 4(b)). A pair of obviously reversible redox peaks in curve a is obtained, suggesting fast electron transfer at the bare gold plate electrode surface. Upon modification with COF_{TFPB-TZT}, the sharply decreased peak current in curve b indicates that the COF_{TFPB-TZT} materials can hamper electron transfer to the electrode surface. When the COF_{TFPB-TZT}-modified electrode is treated with Hg²⁺

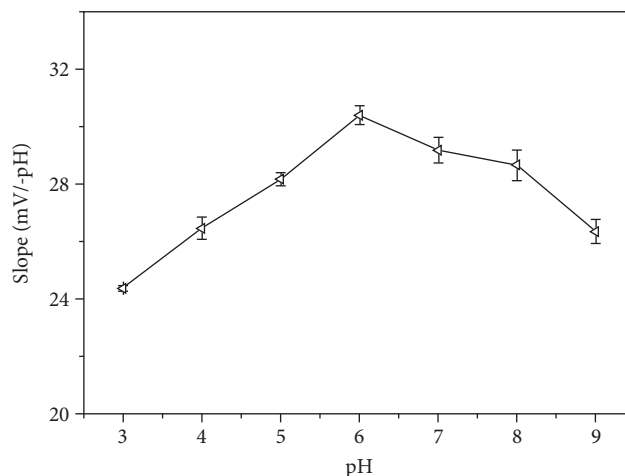


FIGURE 5: Effect of pH on the sensitivity of the GPE/COFTFPB-TZT electrode.

(curve c), the peak currents are slightly enhanced, which can be ascribed to the coordination of Mercury (II). The CV results are consistent with the above EIS results, indicating that this method is feasible for the identification and detection of Mercury ions.

3.3. Optimal pH Selection. In the experiment, the relationship between the electrode potential and the concentration of the measured ionic solution at pH 3.0, 4.0, 5.0, 6.0, 7.0, 8.0, and 9.0 is discussed. The response slope was calculated and the relationship between the slope and pH was determined, as shown in Figure 3. As shown in Figure 5, the electrode presents a Nernst response to Hg²⁺ under acidic conditions (pH = 3.0~6.0). The electrode response slope shows the highest value at pH = 6.0, with a slope value of 30.32 ± 0.2 mV/-PC at 25°C, which is close to the theoretical value of the Nernst response slope of 29.58 mV/-PC [29–31]. The electrode slope at pH 3.0 and 4.0 is

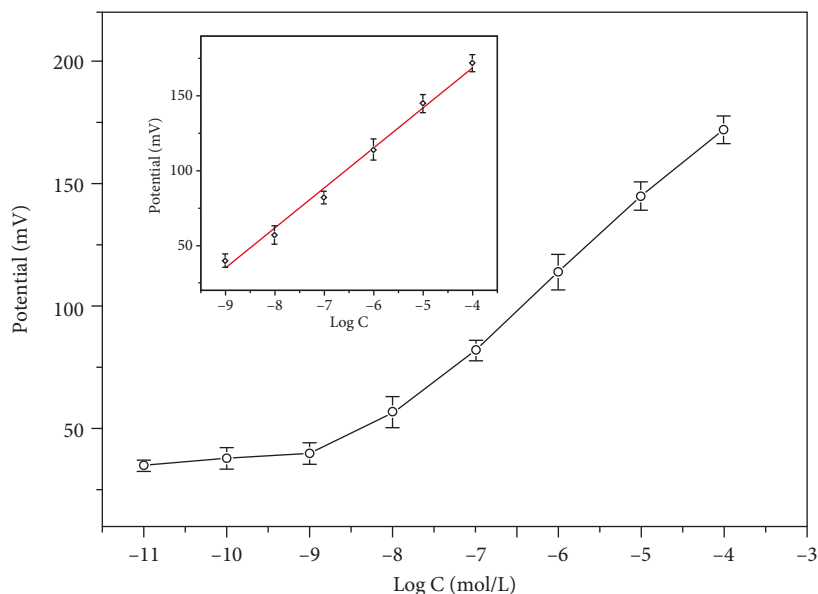


FIGURE 6: Potential response curve for the GPE/COF_{TFPB-TZT} electrode to Hg²⁺ in Tris-HCl buffer (pH = 6.0). The concentrations of Hg²⁺ are $1.0 \times 10^{-9} \sim 1.0 \times 10^{-4}$ mol/L. The inset shows the linear portions of the calibration curve. The error bars are absolute errors deduced from at least three replicate measurements, and the RSDs for all the concentrations are in the range of 3.3~11%.

TABLE 1: Comparison of various nanomaterials based on sensors for Hg²⁺ detection.

No.	Modified electrode	Linear range (μM)	LOD (nM)	Ref.
1	GCE/GA-UiO-66-NH ₂	0.001–2.2	0.9	[32]
2	GCE/L-Cys-Cu ₂ O	0.33–3.3	0.02	[33]
3	GCE/NTOU4-nano-PANI	0.05–27.5	0.035	[34]
4	GCE/ZnFe ₂ O ₄	10–100	80	[35]
5	GCE/ZIF-67/EG	-	1.28	[36]
6	GCE/Cu-MOFs	0.01–0.1	0.0048	[37]
7	GPE/COF _{TFPB-TZT}	0.001–100	0.0045	This work

Note: GCE: glassy carbon electrode; GA: graphene aerogel; NTOU4-nano: zinc phosphite framework nanoparticle; PANI: conductive polyaniline; ZIF-67: cobased zeolitic imidazolate framework; EG: expanded graphite; Cu-MOFs: metal-organic frameworks.

slightly smaller, which can be regarded as the nearly equal near-Nernst response slope. However, the slope of the response gradually decreases as the pH exceeds 6.0, which can be ascribed to the adverse reaction in the solution: $\text{Hg}^{2+} + \text{OH}^- \rightarrow \text{Hg}(\text{OH})^+$. The decrease in slope originates from a small amount of Hg²⁺ being converted to Hg(OH)⁺ for the formation of insoluble salts, leading to a reduction in the concentration of Mercury ions at pH > 6.0. In addition, an insoluble Mercury salt produced in an alkaline solution (pH > 7.0) is likely to adhere to the electrode surface, leading to electrode damage. Therefore, the optimal pH value was selected to be 6.0.

3.4. Electrode Response Performance. The test response performance of GPE/COF_{TFPB-TZT} to Hg²⁺ was investigated in the experiment. Figure 6 shows the potential response curve of the electrode combined with different concentrations of Hg²⁺ in Tris-HCl buffer solution (pH = 6.0). As shown in Figure 6, the electrode potential gradually increases with increasing Hg²⁺ concentration, indicating that the bonding of Hg²⁺ on the modified electrode surface increases. In addition, the electrode shows a good linear response to Hg²⁺ ions in the concentration range of $1.0 \times 10^{-9} \sim 1.0 \times 10^{-4}$ mol/L in Tris-HCl buffer

solution at pH = 6.0. The linear equation fitted by the least square method is $\delta E = 276.08 + 26.82 \log C$ ($R^2 = 0.9873$). The lower limit of detection can be deduced to be 4.5×10^{-12} mol/L according to the plotting method. Comparing the GPE/COF_{TFPB-TZT} with other modified electrodes reported using different nanomaterials [32–37], as shown in Table 1, it can be seen that the COF_{TFPB-TZT}-modified biosensor in this work has better properties with a wider linear range and higher sensitivity.

3.5. Determination of Response Time, Stability, and Reproducibility. The response time and stability of the GPE/COF_{TFPB-TZT}-modified gold plate electrode for Hg²⁺ detection were investigated (Figure 7). The change in the dynamic potential curve after the addition of different concentrations of Hg²⁺ ions to Tris-HCl buffer solution was determined, namely, continuous measurements were carried out from low concentration to high concentration in the range of $1.0 \times 10^{-9} \sim 1.0 \times 10^{-4}$ mol/L and the change in the potential value over time was recorded. The reaction time for the electrode to reach equilibrium in the whole concentration range is very short, which is calculated as 95% of the

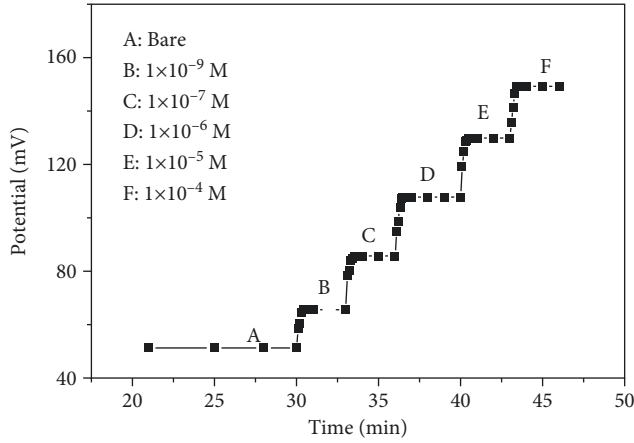


FIGURE 7: Dynamic response of the GPE/COF_{TFPB-TZT} electrode to various concentrations of Hg²⁺ in Tris-HCl buffer (pH = 6.0).

maximum potential response, i.e., 20 s, indicating that the electrode has a fast response speed to Mercury ions. Then, the potential gradually tends to be stable with increasing time, indicating that the response performance of the electrode is stable. At the same time, the electrode was continuously monitored on a 1.0×10^{-7} mol/L Hg²⁺ sample for 25 min with a potential drift of +1.0 mV. The standard deviation for the potential data obtained is ± 2.22 mV, and the relative standard deviation is 1.72% ($n = 10$), indicating that the potential sensor has good stability. After the electrode was tested on Hg²⁺ samples for 15 days, the response slope of the electrode was changed to 20.5 mV/-PC, which decreased by 30.8%, indicating that the potential sensor can be used for at least half a month and has a long service life.

The potential response reproducibility of the GPE/COF_{TFPB-TZT} electrode combined with different concentrations of Hg²⁺ samples was also investigated (Table 2); that is, the potential values of the 1.0×10^{-7} mol/L and 1.0×10^{-4} mol/L Hg²⁺ samples were measured back and forth 10 times. The relative standard deviations are 0.54% and 0.45%, respectively. The relative standard deviation is small, indicating that the electrode has good reproducibility.

3.6. Selectivity of the Electrode. In this experiment, the ion selectivity coefficient of the electrode was measured by the fixed interference ion concentration method (Figure 8), and the selectivity coefficient was calculated by using the Nicolskii-Eisenman formula [18].

$$K_{Hg^{2+}, M^{n+}}^{pot} = \frac{a_{Hg^{2+}}}{a_{M^{n+}}}, \quad (1)$$

where $K_{Hg^{2+}, M^{n+}}^{pot}$ represents the selectivity coefficient of Mercury ions, $a_{Hg^{2+}}$ represents mercury-ion activity, and $a_{M^{n+}}$ represents the activity of interfering ions. In the actual calculation, the ionic strength coefficient is ignored, and the activity is approximately replaced by the concentration. The selectivity coefficients for different metal ions (such as Fe²⁺, Zn²⁺, As³⁺, Cr⁶⁺, Cu²⁺, Cr³⁺, Al³⁺, Pb²⁺, NH₄⁺, Ag⁺, Ba²⁺, Mg²⁺, Na⁺, and K⁺) were investigated. The results show that the selectivity coefficients for these metal ions and oxidizing

TABLE 2: Reproducibility of the GPE/COF_{TFPB-TZT} electrode.

No.	Potential (mV)	
	1.0×10^{-6} mol/L	1.0×10^{-4} mol/L
1	97.0	140.0
2	96.3	139.3
3	97.0	140.0
4	95.6	140.0
5	96.3	139.2
6	96.2	138.5
7	97.1	139.2
8	97.0	139.3
9	96.3	140.0
10	97.0	140.7
AV	96.6	139.6
RSD	0.53%	0.45%

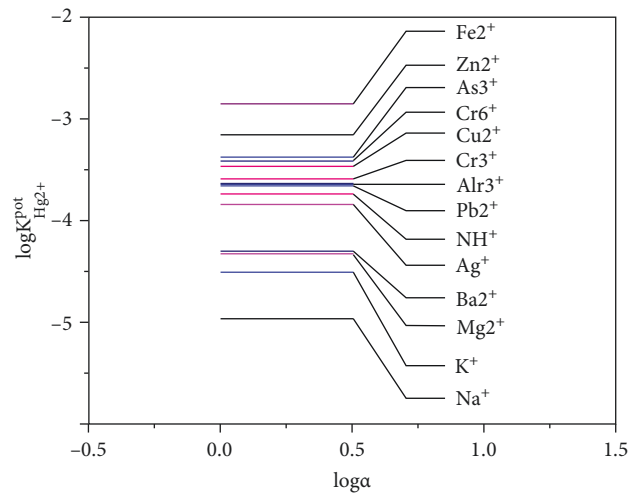


FIGURE 8: Selectivity coefficient of GPE/COF_{TFPB-TZT} to Hg²⁺.

ions are relatively small, which does not interfere with the determination of Hg²⁺ by the electrode, indicating that the electrode shows good selectivity.

3.7. Determination and Analytical Application of the Recovery Rate. To verify the practical applicability of the GPE/COF_{TFPB-TZT} self-assembled gold plate electrode, the fabricated electrodes were used to detect Hg²⁺ in 3 typical water samples (river water, DI water, and polluted water) and compared with the spectrophotometric method with dithizone. As shown in Table 3, the concentration of Hg²⁺ measured by the GPE/COF_{TFPB-TZT} self-assembled gold plate is consistent with that obtained by the spectrophotometric method with dithizone. Moreover, the recovery rates were found to vary in the range from 98.6% to 101.8%. According to the principle of the F test and T test in error analysis, the F value and T value for 6 samples were calculated, respectively, and the F value was determined to be 1.14, 2.14, 1.00, 2.86, 1.15, and 2.12, and the T value was determined to be 1.04, 1.06, 0.64, 0.30, 0.96, and 0.56, respectively. The above-given results indicate that the GPE/COF_{TFPB-TZT} self-assembled gold plate electrode can be used

TABLE 3: Recovery of the proposed GPE/COF_{TFPB-TZT} electrode for the determination of Mercury ions in real water samples compared with the spectrophotometric method utilizing dithizone.

Sample	Spiked	Spectrophotometric method with dithizone		GPE/DTT electrode		Recovery (%)
		($\mu\text{mol/L}$)	SD	($\mu\text{mol/L}$)	SD	
River water	0.500	0.502	± 0.016	0.407	± 0.015	101.8
River water	0.600	0.610	± 0.013	0.491	± 0.019	98.6
DI water	0.700	0.694	± 0.014	0.604	± 0.014	100.7
DI water	0.800	0.794	± 0.013	0.797	± 0.022	99.6
Pollution water	1.100	1.110	± 0.015	1.106	± 0.014	100.5
Pollution water	1.600	1.598	± 0.011	1.596	± 0.016	99.9

to determine Mercury ions in actual water samples. The spectrophotometric method with dithizone is a standard method for the detection of metal ions but requires a complex pretreatment process and involves organic reagents. In comparison with the spectrophotometric method utilizing dithizone, the proposed method is simple, easy to operate, fast, and holds great promise for environmental monitoring, biological medicine, and agricultural product safety testing.

4. Conclusion

A novel N-rich COF_{TFPB-TZT} is prepared with the dehydration condensation reaction of amine aldehydes. A Hg²⁺-selective electrode based on GPE/COF_{TFPB-TZT} is developed based on a self-assembly method. In particular, COF_{TFPB-TZT} contains many N elements, which provide abundant coordination sites for Hg²⁺. The experimental results show that the GPE/COF_{TFPB-TZT} electrode shows excellent performance with a wide linear response range of 1.0×10^{-9} ~ 1.0×10^{-4} mol/L and a short response time of 10 s, and the detection limit is determined to be 4.5 pM. In Tris-HCl buffer solution with pH = 6.0, the electrode potential electrode shows a good potential response to Hg²⁺ that meets the Nernst response. In comparison with the spectrophotometric method utilizing dithizone, the proposed method is simple and fast and holds great potential application prospects in environmental detection and biomedicine.

Data Availability

The data used to support the study are included in the paper.

Consent

Written informed consent for publication was obtained from all authors.

Conflicts of Interest

The authors declare that there are no conflicts of interest.

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