

## Research Article

# Treatment of Distillery Industrial Wastewater Using Ozone Assisted Fenton's Process: Color and Chemical Oxygen Demand Removal with Electrical Energy per Order Evaluation

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Ozonation is one of the most effective and efficient advanced oxidation processes (AOPs) and has shown great potential in the treatment of industrial effluent and wastewater. In the present work, the ozone-Fenton process for % COD and color removal together with electrical energy per order (EE/O) determination for distillery industrial wastewater (DIW) was established. The process was developed by combining the ozone (O<sub>3</sub>) with the Fenton (Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub>) process. The ozone-Fenton (O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub>) was compared with other treatment processes such as O<sub>3</sub>, Fe<sup>2+</sup>, H<sub>2</sub>O<sub>2</sub>, O<sub>3</sub>/Fe<sup>2+</sup>, O<sub>3</sub>/H<sub>2</sub>O<sub>2</sub>, and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> for EE/O together with % COD and color removal efficiency for DIW. The removal of color at 100% and chemical oxygen demand (COD) of 96.875% were achieved with a minimum of EE/O of 0.5315 kWh/m<sup>3</sup> using the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process by operating at optimum conditions. The % COD and color values obtained using O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> were significantly higher than those obtained using O<sub>3</sub>, Fe<sup>2+</sup>, H<sub>2</sub>O<sub>2</sub>, O<sub>3</sub>/Fe<sup>2+</sup>, H<sub>2</sub>O<sub>2</sub> process. The synergy effect of the O<sub>3</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes was evaluated and reported. Our experimental findings suggest that combining O<sub>3</sub> with the Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process could effectively treat industrial effluent and wastewater.

#### 1. Introduction

Nowadays, the ozonation  $(O_3)$  process is commonly used to treat water, wastewater, and industrial effluents in a variety of applications, including disinfection of swimming pool [1], drinking [2], domestic [3], distilleries [4, 5], fermentation [6], dyeing [7], slaughterhouse [8], laundry [9], municipal [10], pharmaceutical [11], synthetic textile [12], RB 5 dye [13], laboratory [14], and petrochemical [15]. The use of  $O_3$ 

in wastewater treatment to disinfect, deodorize, decolorize, and oxidize is becoming essential, particularly when a high degree of treatment is required [6, 16]. It has high efficiency of pollutant removal, an absence of secondary contamination, and a short residence time for wastewater treatment. In comparison to other methods of wastewater treatment such as physical, chemical, biological process, electrochemical [17, 18], and advanced oxidation processes (AOPs) [18, 19], the O<sub>3</sub> process has advantages such as high oxidizing ability,

nonselective simple reaction conditions, without the need for high temperature or pressure [20], etc.

Ozone has a high oxidation potential and produces the hydroxyl radical (<sup>•</sup>OH), a highly reactive oxidative species [21, 22]. While  $O_3$  has high oxidizing power, its conversion to <sup>•</sup>OH is inefficient, and thus O<sub>3</sub> alone has a lower capacity for pollutant removal than OH [23]. Consequently, in recent years, new O3 hybrid processes based on AOPs and electrochemical processes have been developed, including catalytic, nanocatalyzed, photocatalytic, and sonolytic ozonation, O3/UV, O3/H2O2, O3/Fenton, O3/electrochemical, and O<sub>3</sub>/UV/H<sub>2</sub>O<sub>2</sub> [21, 24]. The addition of Fe<sup>2+</sup> and  $H_2O_2$  to  $O_3$ , referred to as the  $O_3/Fe^{2+}/H_2O_2$  process [25], improves the decolorization/degradation/mineralization of organic/inorganic matter present in effluent and wastewater [26, 27]. The O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process produced additional  $^{\circ}$ OH radicals via the reaction of O<sub>3</sub> with Fe<sup>2+</sup>/  $H_2O_2$ , and the reactions are described elsewhere [28, 29].

Owing to the synergic effect, the combination of  $O_3$  with  $Fe^{2+}/H_2O_2$  reagent exhibits a high oxidation rate. The  $O_3/Fe^{2+}/H_2O_2$  process produces excessive °OH via the reaction of  $O_3$  and  $Fe^{2+}/H_2O_2$ , resulting in increased pollutant removal at a reduced treatment time. The  $O_3/Fe^{2+}/H_2O_2$  has been used to treat water present in various types of wastewater, such as urban [30] and landfill leachate [31]. A comparison showed that  $O_3$  and  $Fe^{2+}/H_2O_2$  processes were both less effective than the  $O_3/Fe^{2+}/H_2O_2$  combination [32].

Feng et al. examined the degradation of spent resin using the  $O_3$ -Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes and concluded that both processes run according to first-order kinetic equations, and the  $O_3$ -Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process had the maximum removal competence and cost savings [32]. Li et al. compared the Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> and O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes for amoxicillin degradation. The COD-65% removal rate was achieved in the O<sub>3</sub>-Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process as opposed to the Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process. The synergistic effects worked in harmony with each other, thereby giving the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process the edge [33]. Goi et al. evaluated individual processes such as Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub>, O<sub>3</sub>, and the combination of these treatment processes using leachate waste. They discovered that the coupled Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> and O<sub>3</sub> processes eliminated the most COD (77 %) as compared to other approaches [34].

Substantial research has concentrated on the elimination of pollutants from simulated wastewater using  $O_3$  and AOPs, with only a few studies using real industrial effluent and wastewater. Furthermore, prior research has concentrated on the efficiency of hybrid  $O_3$ -based AOPs in terms of % COD and color removal (%) but has not put an emphasis on pollutant elimination together with electrical energy per order (EE/O). It was crucial to establish the EE/O of  $O_3$ based AOPs to determine the process's operating costs and feasibility. As a result, the current research concentrated on the determination of EE/O while removing color and COD from DIW utilizing the  $O_3$  and  $Fe^{2+}/H_2O_2$  coupled AOPs process.

The primary goal of this study is to assess the efficiency of coupled  $O_3$  and  $Fe^{2+}/H_2O_2$  based AOPs in terms of % COD and color removal, as well as the determination of EE/O in distillery industrial wastewater (DIW). This study investigated

the influence of various operating conditions on the O<sub>3</sub>/Fe<sup>2+</sup>/ $H_2O_2$  process such as Fe<sup>2+</sup> (5–30 mM) and  $H_2O_2$  concentration (20–140 mM), initial pH of wastewater (1–11), COD (800–4800 ppm), and O<sub>3</sub> inlet concentration (0.80–4 g/hr). The synergy between the combined O<sub>3</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes was investigated and recorded.

#### 2. Materials and Methods

2.1. Wastewater Collection and Characterization. Distillery industrial wastewater was collected from distilleries in Erode, Tamil Nadu, India. The wastewater had the following characteristics: dark brown color, burn sugar odor, pH: 4.1–4.3, chemical oxygen demand (COD) of 80,000–90,000 mg/L, biochemical oxygen demand (BOD) of 7000–8000 mg/L, total dissolved solids (TDS) of 5550–5750 mg/L, and total suspended solids (TSS) of 15.44 g/L.

The chemicals used in the experiments were  $H_2O_2-50\%$  (w/w), FeSO<sub>4</sub>.7H<sub>2</sub>O, NaOH,  $H_2SO_4$ , Na<sub>2</sub>S<sub>2</sub>O<sub>3</sub>, KI, K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub>, etc. The analytical reagent (AR) grade chemicals were purchased from Merck, India. They were used as received without any purification.

2.2. Experimental Setup. The experimental setup for the  $O_3/Fe^{2+}/H_2O_2$  process is depicted in Figure 1. The  $O_3$  was generated with the aid of an ozone generator (Ozonetek Limited, Chennai). The air was pumped at a rate of 20 liters per minute (LPM) and generated up to 4 g/hr of  $O_3$ . The generated  $O_3$  is directed into the reactor, which has a capacity of 500 mL. The  $O_3$  was purged via a diffuser at the reactor's bottom. The residual  $O_3$  in the gas stream leaving the reactor was destroyed by the 2–5% KI solution. The DIW was adjusted to the necessary pH and COD concentrations and was loaded into the reactor along with the measured amount of Fe<sup>2+</sup> and H<sub>2</sub>O<sub>2</sub>.

The concentration of  $O_3$  was determined using the iodometric titration process. After the process, the samples were taken from the reactor and immediately added to  $Na_2S_2O_3$  to stop the reaction. After centrifuging for 15 minutes at 15000 rpm, the supernatant was collected and analyzed for color using a UV/Vis–Spectrophotometer (TR300 Spectroquant®) and for COD using the principle of the closed reflux method (TR320, Spectroquant®).

#### 2.3. Analysis

*2.3.1. COD Removal (%).* The COD removal efficiency was measured using the following equation:

COD removal efficiency (%) = 
$$\left(\frac{\text{COD}_{Ini} - \text{COD}_{Fin}}{\text{COD}_{Ini}}\right)$$
100,
(1)

where  $\text{COD}_{Ini}$  and  $\text{COD}_{Fin}$  are the chemical oxygen demand (mg/L) values of DIW before and after the treatment process.



FIGURE 1: Experimental setup of the O<sub>3</sub>/Fe2+/H<sub>2</sub>O<sub>2</sub> process.

2.3.2. Color Removal (%). The color removal efficiency was calculated using the following equation:

color removal (%) = 
$$\left(1 - \frac{Abs_{Fin}}{Abs_{Ini}}\right)$$
100, (2)

where  $Abs_{Ini}$  and  $Abs_{Fin}$  are the absorbances of the before and after treatment process of DIW.

2.3.3. Electrical Energy per Order (EE/O). Electrical energy per order has emerged as an additional and effective way to determine the suitability of wastewater treatment, and it must be economical for both individual and combined processes [35].

The equation proposed for the determination of EE/O for COD removal is as follows:

$$\frac{\text{EE}}{\text{O}} \left( \text{kWh/m}^3 \text{order}^1 \right) = \frac{P * t * 1000}{V * 60 * \log \left( \text{COD}_{\text{Ini}} / \text{COD}_{\text{Fin}} \right)},$$
(3)

$$\log\left(\frac{COD_{Ini}}{COD_{Fin}}\right) = kt,\tag{4}$$

where *P* is the power (kW) for O<sub>3</sub>, t is the treatment time (min), *V* is the volume of the reactor (L), *k* is the pseudo-first-order rate constant (min<sup>-1</sup>) for the deterioration of the pollutant concentration.

Combining equations (3) and (4), EE/O becomes

$$\frac{\text{EP}}{\text{O}}\left(\text{kWh/m}^3 \text{order}^1\right) = \frac{38.4 * P}{V * k}.$$
(5)

2.3.4. Synergistic Effect (SE). The synergy effects (SE) of the  $O_3/Fe^{2+}/H_2O_2$  process can be determined from the COD and/or color removal rate constants of the coupled and standalone processes using the following equation (6) [36]:

$$SE = \left(\frac{k_{O_3/Fe^{2+}/H_2O_2}}{k_{O_3} + k_{Fe^{2+}/H_2O_2}} - 1\right) 100,$$
(6)

where,  $k_{O_3/Fe^{2+}/H_2O_2}$ ,  $k_{O_3}$ ,  $k_{Fe^{2+}/H_2O_2}$  are the rate constants of the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub>, O<sub>3</sub>, and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> system, respectively.

A SE value  $\geq 1$  indicates that the coupled process surpasses the sum of the individuals, instead of SE  $\leq 1$  means that the coupled process produces a negative effect in combining the individuals.

#### 3. Results and Discussion

3.1. Process Comparisons. The individuals such as Fe<sup>2+</sup>,  $H_2O_2$ ,  $O_3$ , and hybrid processes such as  $Fe^{2+}/H_2O_2$ ,  $O_3/Fe^{2+}$ , O<sub>3</sub>/H<sub>2</sub>O<sub>2</sub>, and O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> was carried out under the following operating conditions: COD-1600 mg/L, pH-6,  $Fe^{2+}$ —20 mM,  $H_2O_2$ —100 mM, and  $O_3$  flow rate and concentration of 20 LPM and 4 g/hr, respectively. The efficiency of these processes was compared in terms of % COD and color reduction, accompanied by an estimate of EE/O for DIW, and the findings are depicted in Figures 2(a) and 2(b). As represented in Figure 2(a), the single  $Fe^{2+}$ ,  $H_2O_2$ , and  $O_3$ methods were ineffective at removing color and COD. The Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process removed % COD and color at a moderate rate. The % COD and color removed by combining O3 with  $Fe^{2+}$  and  $H_2O_2$  processes such as  $O_3/Fe^{2+}$ ,  $O_3/H_2O_2$ , and  $O_3/H_2O_2$  $Fe^{2+}/H_2O_2$  were approximately 46.87%, 68.75%, and 96.87%, and 57.43%, 79.74%, and 100%, respectively. As predicted, the hybrid  $O_3/Fe^{2+}/H_2O_2$  process is more efficient at removing COD and color than the  $O_3/Fe^{2+}$  and  $O_3/H_2O_2$ treatment processes. The abovementioned results suggested that adding  $Fe^{2+}$  and  $H_2O_2$  to the  $O_3$  process significantly increases COD and color removal. Perhaps this is due to the influence of concurrent pathways capable of producing plenteous 'OH radicals for COD and color removal from DIW [37, 38].

The suitability of the  $O_3/Fe^{2+}/H_2O_2$  process for wastewater treatment is primarily determined by the EE/O, which relies on COD removal using equation (3). A minimum of  $0.2 \text{ kWh/m}^3 \text{ order}^1$  of EE/O was needed to remove COD and color from DIW using the  $O_3/Fe^{2+}/H_2O_2$  process. In comparison to the  $O_3/Fe^{2+}/H_2O_2$  process, the other single and combined processes such as  $O_3$  and  $O_3/Fe^{2+}$ ,  $O_3/H_2O_2$ required a high amount of EE/O to remove COD and color.

3.2. Various Operating Parameters. Experimental operating parameters including the  $Fe^{2+}$ ,  $H_2O_2$ , COD, and  $O_3$  inlet concentration, and initial wastewater pH [39, 40], and so on, are found to have a major effect on the efficacy of the combined  $O_3/Fe^{2+}/H_2O_2$  process in terms of COD, color removal, and EE/O of DIW.

3.2.1. Effect of  $Fe^{2+}$ . The concentration of  $Fe^{2+}$  and  $H_2O_2$  is an important operating parameter, affecting the efficiency of pollutant removal through the  $O_3/Fe^{2+}/H_2O_2$  process and preventing the excessive use of  $Fe^{2+}$  and  $H_2O_2$  [29, 41]. Figure 3 illustrates the significance of the amount of  $Fe^{2+}$  on the % COD reduction and EE/O in the  $O_3/Fe^{2+}/H_2O_2$ 



FIGURE 2: Comparison of various AOPs such as  $Fe^{2+}$ ,  $H_2O_2$ ,  $Fe^{2+}/H_2O_2$ ,  $O_3$ ,  $O_3/Fe^{2+}$ ,  $O_3/H_2O_3$ , and  $O_3/Fe^{2+}/H_2O_2$  on (a) % color and % COD removal and (b) EE/O (experimental conditions: COD-1600 mg/L; reaction time-4 h; O3 flow rate and production-20 LPM and 4 g/h;  $H_2O_2$ -100 mM;  $Fe^{2+}$ -20 mM; and pH-6).



FIGURE 3: Effect of Fe<sup>2+</sup> on % COD removal and EE/O in the  $O_3/Fe^{2+}/H_2O_2$  process (experimental conditions: COD-1600 mg/L; reaction time-4 h;  $O_3$  flow rate and production-20 LPM and 4 g/h;  $H_2O_2$ -100 mM; and pH-6).

method. As illustrated in Figure 3, around 96.87% COD removal and 0.5315 kWhr/m<sup>3</sup> EE/O were observed in the  $O_3/Fe^{2+}/H_2O_2$  process at a Fe<sup>2+</sup> concentration of 20 mM in comparison to other Fe<sup>2+</sup> concentrations. The combination of Fe<sup>2+</sup> and H<sub>2</sub>O<sub>2</sub> increases the development of °OH and thus the oxidation efficiency. This may be because the increased Fe<sup>2+</sup> concentration facilitated the formation of °OH radicals, thereby accelerating DIW degradation [41]. As Fe<sup>2+</sup> concentrations exceeded 20 mM, the excess Fe<sup>2+</sup> absorbed <sup>•</sup>OH radicals, resulting in a small decline in % COD

reduction and a rise in EE/O of DIW [40]. Thus, the optimal  $Fe^{2+}$  concentration was estimated to be 20 mM.

3.2.2. Effect of  $H_2O_2$ . The influence of varying the  $H_2O_2$  concentration from 20 to 140 mM on the effectiveness of the  $O_3/Fe^{2+}/H_2O_2$  process in terms of % COD reduction and EE/ O for DIW was studied, with the findings shown in Figure 4. The  $H_2O_2$  dose added had a major effect on the studied process. According to Figure 4, the % COD removal



-[]- Electrical energy per order

FIGURE 4: Effect of  $H_2O_2$  on % COD removal and EE/O in the  $O_3/Fe^{2+}/H_2O_2$  process (experimental conditions: COD-1600 mg/L; reaction time-4 h;  $O_3$  flow rate and production-20 LPM and 4 g/h;  $Fe^{2+}$ -20 mM; and pH-6).

increases from 40.63 to 96.87%, and the EE/O decreases from 3.53 to 0.53 kWhr/m<sup>3</sup> as the initial  $H_2O_2$  dose increases to a certain point, reaching a maximum at an initial  $H_2O_2$  dose of about 100 mM, resulting in a substantial increase of process performance. At high concentrations of  $H_2O_2$ , it functions as an efficient <sup>•</sup>OH scavenger [41, 42], depending on the pollutant in concern. The observation is consistent with the following empirical equation (7):

$$H_2O_2 + \bullet OH \longrightarrow HO_2^{\bullet} + H_2O$$
 (7)

$$H_2O_2 \longrightarrow H_2O + \frac{1}{2}O_2.$$
 (8)

While  $HO_2^{\circ}$  promotes radical chain reactions and is an effective oxidant in its own right, it has a much lower oxidation potential than <sup>•</sup>OH. Thus, when H<sub>2</sub>O<sub>2</sub> concentrations are too high, the treatment efficiency is reduced, and its concentration must be tailored for each form of wastewater. The increase in H<sub>2</sub>O<sub>2</sub> dosage increases the number of active sites on the surface, facilitating the decomposition of O<sub>3</sub> molecules into additional <sup>•</sup>OH. Thus, the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process demonstrates an improvement in COD efficiency as the H<sub>2</sub>O<sub>2</sub> dose is increased.

3.2.3. Effect of Wastewater pH. The pH of wastewater at its initial state is critical because it affects the reaction of organic/inorganic compounds with the  $O_3/Fe^{2+}/H_2O_2$  system [25]. Figure 5 shows the results of an investigation into the impact of the initial values of pH on the % COD reduction and EE/O in the  $O_3/Fe^{2+}/H_2O_2$  process for the DIW. As shown in Figure 5, with the  $O_3/Fe^{2+}/H_2O_2$  process, the highest % COD removal efficiency and the lowest EE/O occurred at an initial pH of 6. As the pH increased up to 6, COD removal increased and EE/O decreased. Lower values of pH reveal the H<sup>+</sup> ion's scavenging effect on °OH, the creation of  $H_3O_2^+$  by reaction between  $H_2O_2$  and H<sup>+</sup>, resulting in increased  $H_2O_2$  stability and decreased % COD removal efficiency. The % COD reduction improved as the pH increased from 1 to 6, owing to the increased formation of  $^{\bullet}OH$  (a more powerful oxidant than  $O_3$ ) via the hydroxylation reaction with  $O_3$ .

Furthermore, raising the wastewater pH to 6 in the combined  $O_3/Fe^{2+}/H_2O_2$  process resulted in coagulation, which increases contaminants through the complexation reaction induced by the transformation of  $Fe^{2+}$  and  $Fe^{3+}$  to  $Fe(OH)_n$ -type structures [25]. When the pH was greater than 6, COD removal decreased marginally with increasing pH. It is deduced that COD removal exists as a eupterotid that is energetic and easily reacts with the hydroxyl ion in acidic conditions, but it acts as a stable molecule in basic conditions, resulting in a decrease in % COD reduction and an increase in EE/O as the pH increases.

3.2.4. Effect of COD Concentration. The initial pollutant content is critical in the  $O_3/Fe^{2+}/H_2O_2$  process wastewater treatment process [43, 44]. The increased initial COD concentration from 800 to 4800 mg/L, then decreased the % COD elimination from 100 to 45.83%, with a rise in the EE/O from 0.07 to 3 kWh/m<sup>3</sup>order<sup>1</sup> for DIW using the  $O_3/Fe^{2+}/H_2O_2$  process, as presented in Figure 6. Since, raising the initial COD content promoted the creation of intermediate products that compete with the  $O_3$ -consuming pollutant, thus decreasing % COD removal performance, and increase of EE/O for DIW [45].

3.2.5. Effect of  $O_3$  Inlet Concentration. It is critical to select the optimal  $O_3$  inlet concentration for pollutant removal from wastewater when using the  $O_3/Fe^{2+}/H_2O_2$  process [40, 46]. The effect of  $O_3$  inlet concentration on the % COD removal efficiency and EE/O with the  $O_3/Fe^{2+}/H_2O_2$  process was explored by bubbling ozone into the DIW solution at different gas concentrations ranging from 0.8 to 4 g/h. As



FIGURE 5: Effect of pH on % COD removal and EE/O in the  $O_3/Fe^{2+}/H_2O_2$  process (experimental conditions: COD-1600 mg/L; reaction time-4 h;  $O_3$  flow rate and production-20 LPM and 4 g/h;  $H^2O^2$ -100 mM;  $Fe^{2+}$ -20 mM; and pH-6).



FIGURE 6: Effect of COD on % COD removal and EE/O in the  $O_3/Fe^{2+}/H_2O_2$  process (experimental conditions: reaction time-4 h;  $O_3$  flow rate and production-20 LPM and 4 g/h;  $H_2O_2$ -100 mM;  $Fe^{2+}$ -20 mM; and pH-6).

shown in Figure 7, the % COD reduction improved from 52.18 to 96.87% and the EE/O reduced from 2.49 to  $0.535 \text{ kWh/m}^3$  order<sup>1</sup> as the  $O_3/\text{Fe}^{2+}/\text{H}_2O_2$  system's  $O_3$  concentration increased from 0.8 to 4 g/hr. This can be explained by the fact that the two key parameters,  $O_3$  concentration and  $O_3$ -liquid mass transfer resistance, have a massive effect on the mass transfer rate of  $O_3$  [40, 46]. With a higher  $O_3$  concentration, the driving factor for  $O_3$  mass transfer is increased, allowing the DIW solution to absorb more  $O_3$ . As a consequence, the excess  $O_3$  in the solution interacted with radical initiators (Fe<sup>2+</sup>, H<sub>2</sub>O<sub>2</sub>, °OH, etc.) to produce more °OH, which eventually improves COD and color removal efficiency while lowering EE/O. Zhao et al.

[46] reported similar results for the removal of Ni-EDTA using O<sub>3</sub>-based oxidation processes.

3.3. Synergy Effect. The O<sub>3</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes were carried out under the best experimental conditions to determine the synergy among each process for the removal of % color and % COD from DIW. The experimental findings confirmed a synergy index among O<sub>3</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> processes for COD removal of DIW. Thus, as opposed to other processes (O<sub>3</sub> and Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub>), the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process greatly improves COD and color removal. Equation (6) was used to determine the synergy index of the O<sub>3</sub> and Fe<sup>2+</sup>/



FIGURE 7: Effect of O<sub>3</sub> inlet concentration on % COD removal and EE/O in the O<sub>3</sub>/Fe<sup>2+</sup>/H<sub>2</sub>O<sub>2</sub> process (experimental conditions: COD-1600 mg/L; reaction time-4 h; O<sub>3</sub> flow rate-20 LPM; H2O2-100 mM; Fe<sup>2+</sup>-20 mM; and pH-6).

 $H_2O_2$  processes [36]. The synergy index was 21.50%, suggesting that the effectiveness in terms % color and % COD removal was sufficiently greater for the combined process than for the individual  $O_3$  and  $Fe^{2+}/H_2O_2$  processes. The reported synergy index is due to the fact that combining  $O_3$  and  $Fe^{2+}/H_2O_2$  provides a larger volume of °OH, which enhances the rate of % color and COD removal from DIW. Thus, the coupled  $O_3$  and  $Fe^{2+}/H_2O_2$  processes offer an alternative and novel approach for industrial wastewater treatment.

#### 4. Conclusion

The developed  $O_3/Fe^{2+}/H_2O_2$  process is compared to  $O_3$ ,  $Fe^{2+}$ ,  $H_2O_2$ ,  $O_3/Fe^{2+}$ , and  $O_3/H_2O_2$ ,  $Fe^{2+}/H_2O_2$  processes, and the outcomes indicate that the  $O_3/Fe^{2+}/H_2O_2$  process is a promising method for achieving high % color and % COD removal efficiencies while consuming a minimal amount of EE/O from DIW. The findings suggest that to ensure an effective treatment process, operating parameters such as  $Fe^{2+}$ ,  $H_2O_2$ , COD, and  $O_3$  inlet concentration, and initial wastewater pH should be considered. A synergy effect was calculated to exist between the  $O_3$  and  $Fe^{2+}/H_2O_2$  processes at 21.50 percent. The  $O_3/Fe^{2+}/H_2O_2$  method was found to be capable of eliminating pollutants from a range of industrial effluents and wastewater.

#### **Data Availability**

The datasets analyzed during the study are available from the corresponding author on request.

### **Conflicts of Interest**

The authors acknowledge that they have no conflicts of interest.

#### References

- K. M. S. Hansen, A. Spiliotopoulou, W. A. Cheema, and H. R. Andersen, "Effect of ozonation of swimming pool water on formation of volatile disinfection by-products—a laboratory study," *Chemical Engineering Journal*, vol. 289, pp. 277–285, 2016.
- [2] Y. Li, W. Shen, S. Fu, H. Yang, G. Yu, and Y. Wang, "Inhibition of bromate formation during drinking water treatment by adapting ozonation to electro-peroxone process," *Chemical Engineering Journal*, vol. 264, pp. 322–328, 2015.
- [3] Z. A. Yacouba, J. Mendret, G. Lesage, F. Zaviska, and S. Brosillon, "Removal of organic micropollutants from domestic wastewater: the effect of ozone-based advanced oxidation process on nanofiltration," *Journal of Water Process Engineering*, vol. 39, Article ID 101869, 2021.
- [4] S. N. Malik, P. C. Ghosh, A. N. Vaidya, and S. N. Mudliar, "Ozone pretreatment of biomethanated distillery wastewater in a semi batch reactor: mapping pretreatment efficiency in terms of COD, color, toxicity and biohydrogen generation," *Biofuels*, vol. 11, pp. 801–809, 2020.
- [5] B. Otieno, S. Apollo, B. Naidoo, and A. Ochieng, "Modeling ozonation pretreatment parameters of distillery wastewater for improved biodegradability," *Journal of Environmental Science and Health—Part A Toxic/Hazardous Substances and Environmental Engineering*, vol. 54, no. 11, pp. 1066–1074, 2019.
- [6] E. Rahimi, Y. D. Shahamat, B. Kamarehei, A. Zafarzadeh, and M. R. Khani, "Rapid decolorization and mineralization of molasses by catalytic ozonation process with a nanocomposite from fermentation industry wastewater," *International Journal of Environmental Science and Technology*, vol. 15, no. 9, pp. 1941–1948, 2018.
- [7] X.-L. Zou, "Combination of ozonation, activated carbon, and biological aerated filter for advanced treatment of dyeing wastewater for reuse," *Environmental Science and Pollution Research*, vol. 22, no. 11, pp. 8174–8181, 2015.
- [8] P. Alfonso-Muniozguren, A. I. Gomes, D. Saroj, V. J. P. Vilar, and J. Lee, "The role of ozone combined with UVC/H2O2

process for the tertiary treatment of a real slaughterhouse wastewater," *Journal of Environmental Management*, vol. 289, Article ID 112480, 2021.

- [9] V. V. Patil, P. R. Gogate, A. P. Bhat, and P. K. Ghosh, "Treatment of laundry wastewater containing residual surfactants using combined approaches based on ozone, catalyst and cavitation," *Separation and Purification Technology*, vol. 239, Article ID 116594, 2020.
- [10] S. Dogruel, Z. Cetinkaya Atesci, E. Aydin, and E. Pehlivanoglu-Mantas, "Ozonation in advanced treatment of secondary municipal wastewater effluents for the removal of micropollutants," *Environmental Science and Pollution Research*, vol. 27, no. 36, pp. 45460–45475, 2020.
- [11] Y. Shi, S. Li, L. Wang et al., "Compositional characteristics of dissolved organic matter in pharmaceutical wastewater effluent during ozonation," *The Science of the Total Environment*, vol. 778, Article ID 146278, 2021.
- [12] Y. C. Yang, S. S. Zeng, Y. Ouyang et al., "An intensified ozonation system in a tank reactor with foam block stirrer: synthetic textile wastewater treatment and mass transfer modeling," *Separation and Purification Technology*, vol. 257, Article ID 117909, 2021.
- [13] N. P. Chokshi and J. P. Ruparelia, "Synthesis of nano Ag-La-Co composite metal oxide for degradation of RB 5 dye using catalytic ozonation process," *Ozone: Science and Engineering*, vol. 44, no. 2, pp. 182–195, 2021.
- [14] I. Basturk, S. Murat-Hocaoglu, G. Varank, and S. Yazici-Guvenc, "Comparison of ozonation and electro-Fenton processes for Sodium Azide removal in medical laboratory wastewater by using central composite design," *Separation Science and Technology*, vol. 56, 2021.
- [15] L. Jothinathan, Q. Q. Cai, S. L. Ong, and J. Y. Hu, "Organics removal in high strength petrochemical wastewater with combined microbubble-catalytic ozonation process," *Chemosphere*, vol. 263, Article ID 127980, 2021.
- [16] T. Watanabe, M. Tanaka, K. Masaki, T. Fujii, and H. Iefuji, "Decolorization and treatment of Kokuto-shochu distillery wastewater by the combination treatment involving biodecolorization and biotreatment by Penicillium oxalicum d, physical decolorization by ozonation and treatment by activated sludge," *Biodegradation*, vol. 21, no. 6, pp. 1067–1075, 2010.
- [17] M. Dolatabadi, T. Świergosz, and S. Ahmadzadeh, "Electro-Fenton approach in oxidative degradation of dimethyl phthalate—the treatment of aqueous leachate from landfills," *The Science of the Total Environment*, vol. 772, Article ID 145323, 2021.
- [18] S. Fekadu, E. Alemayehu, P. Asaithambi, and B. Van der Bruggen, "Removal of phosphate from the healthcare wastewater through peroxi-photoelectrocoagulation process: effect of process parameters," *International Journal of Environmental Research*, vol. 16, no. 1, p. 8, 2022.
- [19] M. Dolatabadi, M. T. Ghaneian, C. Wang, and S. Ahmadzadeh, "Electro-Fenton approach for highly efficient degradation of the herbicide 2,4-dichlorophenoxyacetic acid from agricultural wastewater: process optimization, kinetic and mechanism," *Journal of Molecular Liquids*, vol. 334, Article ID 116116, 2021.
- [20] P. Alfonso-Muniozguren, M. Hazzwan Bohari, A. Sicilia et al., "Tertiary treatment of real abattoir wastewater using combined acoustic cavitation and ozonation," *Ultrasonics Sonochemistry*, vol. 64, Article ID 104986, 2020.
- [21] S. N. Malik, P. C. Ghosh, A. N. Vaidya, and S. N. Mudliar, "Hybrid ozonation process for industrial wastewater

treatment: principles and applications: a review," *Journal of Water Process Engineering*, vol. 35, Article ID 101193, 2020.

- [22] Y. Yoon, Y. Hwang, M. Kwon, Y. Jung, T.-M. Hwang, and J.-W. Kang, "Application of O3 and O3/H2O2 as posttreatment processes for color removal in swine wastewater from a membrane filtration system," *Journal of Industrial and Engineering Chemistry*, vol. 20, no. 5, pp. 2801–2805, 2014.
- [23] Z. Yang, Y. Zhang, W. Zhu, X. Zan, L. Zhang, and Y. Liu, "Effective oxidative degradation of coal gasification wastewater by ozonation: a process study," *Chemosphere*, vol. 255, Article ID 126963, 2020.
- [24] C. V. Rekhate and J. K. Srivastava, "Recent advances in ozonebased advanced oxidation processes for treatment of wastewater-a review," *Chemical Engineering Journal Advances*, vol. 3, Article ID 100031, 2020.
- [25] S. S. Abu Amr and H. A. Aziz, "New treatment of stabilized leachate by ozone/Fenton in the advanced oxidation process," *Waste Management*, vol. 32, no. 9, pp. 1693–1698, 2012.
- [26] J. Wang and H. Chen, "Catalytic ozonation for water and wastewater treatment: recent advances and perspective," *Science of the Total Environment*, vol. 704, Article ID 135249, 2020.
- [27] S. Pirgalioğlu and T. A. Özbelge, "Comparison of non-catalytic and catalytic ozonation processes of three different aqueous single dye solutions with respect to powder copper sulfide catalyst," *Applied Catalysis A: General*, vol. 363, pp. 157–163, 2009.
- [28] P. Asaithambi, B. Sajjadi, A. R. Abdul Aziz, and W. M. A. B. W. Daud, "Ozone (O3) and sono (US) based advanced oxidation processes for the removal of color, COD and determination of electrical energy from landfill leachate," *Separation and Purification Technology*, vol. 172, pp. 442–449, 2017.
- [29] C. V. Rekhate and J. K. Srivastava, "Effectiveness of O3/Fe2+/ H2O2 process for detoxification of heavy metals in municipal wastewater by using RSM," *Chemical Engineering and Processing*, vol. 165, Article ID 108442, 2021.
- [30] J. Mendret, A. Azais, T. Favier, and S. Brosillon, "Urban wastewater reuse using a coupling between nanofiltration and ozonation: techno-economic assessment," *Chemical Engineering Research and Design*, vol. 145, pp. 19–28, 2019.
- [31] C. Wu, W. Chen, Z. Gu, and Q. Li, "A review of the characteristics of fenton and ozonation systems in landfill leachate treatment," *Science of the Total Environment*, vol. 762, Article ID 143131, 2021.
- [32] W. Feng, J. Li, K. Gao, H. An, and Y. Wang, "A comparison of spent resin degradation by Fenton and O3-Fenton process," *Progress in Nuclear Energy*, vol. 130, Article ID 103566, 2020.
- [33] M. Li, Z. Zeng, Y. Li et al., "Treatment of amoxicillin by O 3/ Fenton process in a rotating packed bed," *Journal of Envi*ronmental Management, vol. 150, pp. 404–411, 2015.
- [34] A. Goi, Y. Veressinina, and M. Trapido, "Combination of ozonation and the Fenton processes for landfill leachate treatment: evaluation of treatment efficiency," *Ozone: Science* & Engineering, vol. 31, no. 1, pp. 28–36, 2009.
- [35] C.-H. Wu and H.-Y. Ng, "Degradation of C.I. Reactive Red 2 (RR2) using ozone-based systems: comparisons of decolorization efficiency and power consumption," *Journal of Hazardous Materials*, vol. 152, no. 1, pp. 120–127, 2008.
- [36] J. Madhavan, F. Grieser, and M. Ashokkumar, "Degradation of orange-G by advanced oxidation processes," *Ultrasonics Sonochemistry*, vol. 17, no. 2, pp. 338–343, 2010.
- [37] Q. Wei, S. Qiao, B. Sun, H. Zou, J. Chen, and L. Shao, "Study on the treatment of simulated coking wastewater by O3 and

O3/Fenton processes in a rotating packed bed," RSC Advances, vol. 5, no. 113, pp. 93386–93393, 2015.

- [38] L. Mansouri, M. Sabelfeld, S.-U. Geissen, and L. Bousselmi, "Catalysed ozonation for removal of an endocrine-disrupting compound using the O3/Fenton reagents system," *Environmental Technology*, vol. 36, no. 13, pp. 1721–1730, 2015.
- [39] A. Ikhlaq, H. M. S. Munir, A. Khan, F. Javed, and K. S. Joya, "Comparative study of catalytic ozonation and Fenton-like processes using iron-loaded rice husk ash as catalyst for the removal of methylene blue in wastewater," *Ozone: Science & Engineering*, vol. 41, no. 3, pp. 250–260, 2019.
- [40] M. K. Nguyen, V. S. Tran, T. T. Pham et al., "Fenton/ozonebased oxidation and coagulation processes for removing metals (Cu, Ni)-EDTA from plating wastewater," *Journal of Water Process Engineering*, vol. 39, Article ID 101836, 2021.
- [41] C. Kalyanaraman, K. S. B. Kameswari, and J. R. Rao, "Studies on enhancing the biodegradation of tannins by ozonation and Fenton's oxidation process," *Journal of Industrial and Engineering Chemistry*, vol. 25, pp. 329–337, 2015.
- [42] L. Prieto-Rodríguez, D. Spasiano, I. Oller, I. Fernández-Calderero, A. Agüera, and S. Malato, "Solar photo-Fenton optimization for the treatment of MWTP effluents containing emerging contaminants," *Catal Today*, vol. 209, pp. 188–194, 2013.
- [43] M. Sui, S. Xing, L. Sheng, S. Huang, and H. Guo, "Heterogeneous catalytic ozonation of ciprofloxacin in water with carbon nanotube supported manganese oxides as catalyst," *Journal of Hazardous Materials*, vol. 227–228, pp. 227–236, 2012.
- [44] G. Moussavi and M. Mahmoudi, "Degradation and biodegradability improvement of the reactive red 198 azo dye using catalytic ozonation with MgO nanocrystals," *Chemical En*gineering Journal, vol. 152, no. 1, pp. 1–7, 2009.
- [45] H. T. Van, L. H. Nguyen, T. K. Hoang et al., "Using FeOconstituted iron slag wastes as heterogeneous catalyst for Fenton and ozonation processes to degrade reactive red 24 from aqueous solution," *Separation and Purification Technology*, vol. 224, pp. 431–442, 2019.
- [46] Z. Zhao, Z. Liu, H. Wang, W. Dong, and W. Wang, "Sequential application of Fenton and ozone-based oxidation process for the abatement of Ni-EDTA containing nickel plating effluents," *Chemosphere*, vol. 202, pp. 238–245, 2018.