Research Article

UV-Irradiated Photocatalytic Degradation of Nitrobenzene by Titania Binding on Quartz Tube

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A new method for UV-irradiated degradation of nitrobenzene by titania photocatalysts was proposed, titania nanoparticles were coated on a quartz tube through the introduction of tetraethyl orthosilicate into the matrix. The dependence of nitrobenzene photodegradation on pH, temperature, concentration, and air feeding was discussed, and the physical properties such as the activation energy, entropy, enthalpy, adsorption constant, and rate constant were acquired by conducting the reactions in a variety of experimental conditions. The optimum efficiency of the photodegradation with the nitrobenzene residue as low as 8.8% was achieved according to the experimental conditions indicated. The photodegradation pathways were also investigated through HPLC, GC/MS, ion chromatography (IC), and chemical oxygen demand (COD) analyses.

1. Introduction

Wastewater discharged from industrial effluents has been a serious problem of the environmental issues [1]. Due to the relatively stable chemical structures of aromatic dyes, most of the dyestuffs that the industries produced are unlikely to decompose and therefore causing severe contaminations to the drinking water and irrigation systems. Many of the dyestuffs have even been considered health hazards and thus serious concerns have been raised [2, 3]. Nitrobenzene (NB) is one of the frequently used chemicals for aniline manufacturing, rubber making, textile, pesticide, cosmetic, and pharmaceutical purposes [4]. It is a highly toxic pollutant. Through skin contact and inhalation, it is readily absorbed into human body and is likely causing symptoms such as headache, vertigo, vomit, and weakness. For this reason, it was selected to be the model pollutant in this study. General treatments of such polluted wastewater include adsorption [5, 6], ozonation [7], or biodegradation [8]. However, those methods are somewhat inefficient or relatively expensive because they require further treatments and are very prone to cause secondary pollution.

Semiconductor photocatalysts have been considered potential candidates for treating various water pollutants [9-11] and are acclaimed as being one of advanced oxidation processes for eliminating the harmful contaminants. The efficient oxidation reaction of photocatalysts under UV irradiation processes provides an alternative way of managing these pollutants properly. Due to the extensive research for many years, semiconductor photocatalysts which have been frequently investigated include oxide compounds of TiO₂, ZnO, WO₃, SnO₂, ZrO₂, and sulfide compounds of CdS and ZnS [12, 13]. Among these compounds, titania (TiO₂) has been thoroughly investigated for its photocatalytic activities [14-18]. Novel preparations of titania with high photocatalytic activities have been proposed [19, 20]. It has drawn great attention in research and industries in recent years owing to its powerful oxidation capability, non-toxicity, biological inertness, chemical stability, and costeffectiveness [21]. Recently, selective photocatalysis of titania has been developed and gained success, it can be applied to separation process or selective elimination of specific molecules from a mixture [22, 23]. Some research explored further the photocatalytic activities of mixed oxides [24], and some studied the effect of morphologies of titania [25, 26]. In addition, titania photocatalysis has been applied to solar disinfection of water [27] and degrading chemicals that caused hormonal imbalance in wildlife and humans [28].

The initial step involves the generation of e^-/h^+ pairs after irradiated with light of which the wavelength is shorter than 388 nm (413 nm for rutile crystalline phase), due to the wide direct band gap of TiO₂ (3.2, 3.0 eV for anatase and rutile crystalline phase, resp.), the valence band electrons of TiO₂ will be promoted up to the conduction band, leaving holes in valence band as shown in (1). The holes produced could subsequently react with H₂O molecules or hydroxide ions (OH⁻) in aqueous solution forming powerful oxidants such as hydroxyl radicals (•OH) in (2). The electrons on conduction band could transform the dissolved oxygen to superoxide radicals in (3) and further to hydrogen peroxide in (4)–(7) [29]. These strong oxidants could be used to mineralize organic pollutants from wastewater as follows:

$$\text{TiO}_2 + h\nu \longrightarrow \text{TiO}_2(e^- + h^+)$$
 (1)

$$h^+ + H_2O \longrightarrow OH + H^+(h^+ + OH^- \longrightarrow OH$$

in alkaline condition)

(2)

$$e^- + O_2 \longrightarrow O_2^{\bullet -} \tag{3}$$

$$O_2^{\bullet-} + H^+ \longrightarrow HO_2^{\bullet}$$
 (4)

$$\mathrm{HO}_{2}^{\bullet} + \mathrm{HO}_{2}^{\bullet} \longrightarrow \mathrm{H}_{2}\mathrm{O}_{2} + \mathrm{O}_{2} \tag{5}$$

$$O_2^{\bullet-} + HO_2^{\bullet} \longrightarrow HO_2^- + O_2$$
 (6)

$$\mathrm{HO}_{2}^{-} + \mathrm{H}^{+} \longrightarrow \mathrm{H}_{2}\mathrm{O}_{2}. \tag{7}$$

The photocatalytical degradation of NB may undergo either through direct oxidation reduction from electrons and holes or through indirect reaction with oxidants such as hydroxyl radicals, superoxide radicals, and hydrogen peroxide, which are generated by UV irradiation onto TiO₂.

The aim of this study is to investigate the optimized experimental conditions for the photodegradation of NB and to propose a new approach which can easily be applied repetitively without the necessity of separating the catalyst from solution after use. By employing HPLC, GC/MS, ion chromatography (IC), and chemical oxygen demand (COD) analyses, the authors attempted to probe into the photodegradation pathways. The structural and optical characterization of TiO₂ has been done by means of X-ray diffraction (XRD) and UV-vis spectroscopy (UV-vis). The photocatalytic activities of TiO₂ nanoparticles were evaluated through the degradations of NB solutions initiated by irradiation with an UV lamp.

2. Experimental

2.1. Materials. Degussa P25 titania (TiO₂, 80% in anatase phase, average particle size 30 nm, surface area $50 \text{ m}^2 \text{ g}^{-1}$) was acquired from Degussa Co. Nitrobenzene was purchased from Panreac Co., *m*-, *o*-, *p*-nitrophenol, tetraethyl orthosilicate, perchloric acid, HPLC grade methanol, sodium nitrate,

sodium nitrite, sodium carbonate, and sodium bicarbonate were acquired from Merck Co. Ethanol was purchased from Riedel-de Haen Co. Coumarin was purchased from Alfa Aesar Co. These chemicals were used without further purification. Ultrapure water ($18 M\Omega$ -cm) was used throughout the entire experiments.

2.2. Preparation of Tetraethyl Orthosilicate Polymer Solution. One mole of tetraethyl orthosilicate and 2.5 moL ethanol were mixed for 30 min at room temperature, hydrochloric acid solution was added to make the solution with a pH value of pH 1. Then a mixture of 2.5 moL ethanol and 4 moL H₂O was added into the solution above for 30 min to form a polymer solution which is in molar ratio of 1:4:5 for tetraethyl orthosilicate, H₂O, and ethanol. A quartz tube (2.2 cm in outer diameter and 28 cm in length) was first dipped into the solution composed of TiO₂ and polymer (1:20 w/w ratio) three times and was sintered in an oven at 400°C for 1 h.

2.3. NB Photodegradation Reactions. A 482 mL stainless steel tank was used as the container for TiO₂ photocatalytic degradation of NB solutions, inside the container the water bath was circulated by a peristaltic pump at a flow rate of 73 mL min⁻¹ and thermostated by a HAAKE DC10 immersion circulator at 30°C. The quartz tube referred to above was put inside a glass tube, which is 5 cm in inner diameter and 30 cm in length and was used as the batch reactor. A 14 W Philips UV lamp with the wavelength of 254 nm was inserted into the quartz tube. Silicone tubing was introduced inside the reactor for feeding air into the solution. The degradation experiments were carried out by mixing each 400 mL of 0.407, 0.813, 1.220, 1.626, and 2.439 mM NB solution with 0.2 g TiO₂ photocatalyst supported on the quartz tube and covered with aluminum foil, the selected air feeding was 50, 100, and 150 mL min⁻¹ (the dissolved oxygen is 5.8, 6.5, and 7.2 mg L^{-1} , resp.), the pH values of solutions were adjusted to pH 4, 7, and 11 using 0.1 N perchloric acid or 0.1 N NaOH solution. The solutions were sampled at 0, 15, 30, 45, 60, 90, 120, 150, and 180 min reaction time for HPLC, IC, and GC/MS analyses. The absorption maximum of the UV-vis absorbance spectroscopy appeared at 270 nm (λ_{max}) for NB solutions and was selected to be the analytical wavelength for NB measurements. For the identifications of intermediates, the photodegradation was manipulated in a slurry system with 1.0 g catalyst and 1.626 mM NB at pH 7 and 30° C with the air feeding of 150 mL min⁻¹ for 10 h.

2.4. Characterization of TiO_2 Nanoparticles and Quantitative Analyses of Reaction Intermediates. A Shimadzu XD-D1 Xray diffractometer was used to observe the XRD patterns of nanoparticles, and the 2θ range of 20° to 70° was scanned continuously with a step of 0.02° and at a speed of 1 degree min⁻¹. A Hitachi U-2000 UV-vis spectrophotometer was used to evaluate the degradation of NB solutions, the absorption maximum (λ_{max}) of the UV-vis absorbance spectroscopy of NB solutions was selected at 270 nm based on analytical sensitivity. A Hitachi L-7000 high-performance liquid chromatography (HPLC; Hitachi L-7100 pump) using a C18 column (Waters $5 \mu m$, 250 mm long) connected with a Hitachi L-7420 UV detector was used for the qualitative and quantitative determinations of the initial compound and reaction intermediates. A Dionex DX-120 ion chromatography (IC) system using IonPac AS9-HC analytical column (4 \times 250 mm, 1700–1800 psi) and IonPac AG9-HC guard column (4 \times 50 mm) with 9 mM Na₂CO₃ as the mobile phase was used for the qualitative and quantitative determinations of anions produced during the reaction. A Finnigan Trace MS connected with Trace GC 2000 GC/MS in electron impact (EI) mode were used to separate and analyze the intermediates of reaction. Samples were injected splitless into the gas chromatography equipped with a Rtx-5 ms column (30 m \times 0.25 mm, 0.25 μ m film thickness). The oven was programmed isotherm at 35°C for 2 min, then heating from 35°C to 250°C at the speed of 10°C min⁻¹, and keep isotherm at 250°C for 1 min. An AQUALYTIC AL31 chemical oxygen demand reactor was used to detect the amount of organic compound residue in solution.

2.5. Measurements of Hydroxyl Radicals Generated from Titania under UV Irradiation. The measurement of hydroxyl radicals was proposed using coumarin or terephthalic acid as probe molecules [30, 31]. Coumarin was selected as the probe molecule in this work. Each 0.1 g titania powder was dispersed in 20 mL 1.0 mM coumarin aqueous solution until it reached adsorption-desorption equilibrium before light irradiation. In a dish with 7.0 cm in diameter, the solution was located 10 cm below the light source and was irradiated using an UV lamp with the peak intensity of 366 nm. After UV illumination the solution was centrifuged and the PL spectrum of hydroxylated coumarin (7-hydroxycoumarin) was measured by a Hitachi F-4500 fluorescence spectrophotometer with the excitation wavelength of 332 nm, scanning speed 240 nm/min, PMT voltage 700 V, and slit width 2.5 nm. To investigate the effect of pH value on the production of hydroxyl radicals, the coumarin solution was adjusted to pH 4, 7, and 11 using 0.1 N perchloric acid or 0.1 N NaOH solution.

3. Results and Discussion

3.1. Background Experiments. To explore the background interferences, volatility of solvent, adsorption, and blank experiments were carried out to confirm the role of titania in the photocatalysis reactions. The results of volatility experiments showed no significant decrease of NB concentration was observed after 3 h. With respect to the adsorption of NB by the catalysts, the adsorption tests showed that saturated adsorption of NB was reached after about 150 min of adsorption. The results of blank experiments (direct photocatalysis of NB solution without the catalysts) showed more than 71.1% NB remained after 3 h irradiation under a variety of conditions. Therefore, the concentrations of NB were examined after 3 h of adsorption for all the degradation experiments.

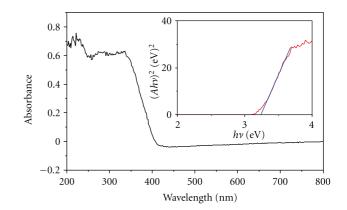


FIGURE 1: The UV-vis absorbance spectrum of TiO₂ photocatalyst. Inset shows the plot of $(Ahv)^2$ versus hv for TiO₂, the straight line represents the extrapolation of the abscissa from the linear fit of the curve, through which the band gap was estimated at about 3.20 eV.

3.2. Characterization of TiO_2 Photocatalysts. From the UVvis spectrum in Figure 1, very little light could be absorbed in the spectrum with wavelength greater than 420 nm by semiconductor TiO₂. The energy gap of TiO₂ nanoparticles can be obtained from the data treatment of optical absorbance by plotting $(Ahv)^2$ versus hv according to the following relationship [32, 33]:

$$A = \frac{\left[k\left(hv - E_g\right)^{n/2}\right]}{hv},\tag{8}$$

where A is the absorbance, k is a constant, h is the Planck's constant, v is the frequency of light, n = 1 for direct and n = 4 for indirect electron transfer, and E_g is the band gap of the object concerned. From the inset of Figure 1, P25 TiO₂ was estimated having the energy gap of 3.2 eV, which is in accordance with other research [34]. From the XRD patterns of TiO₂ photocatalysts (not shown), the characteristic signals of anatase crystalline phase TiO₂ were found at $2\theta = 25.4^{\circ}$ (1) 0 1), 37.9° (0 0 4), 48.2° (2 0 0), 55.1° (2 1 1), 62.8° (2 0 4), and the signals for rutile crystalline phase at 27.5° (1 1 0), 36.2° (1 0 1), 54.4° (2 1 1). Using the Scherrer equation with a shape factor of 0.89, $\lambda = 0.1541$ nm for the wavelength of the Cu K α_1 X-ray source, and B the full peak width at half-maximum, the particle size (d; in nm) of nanoparticles can be estimated [35]. Applying XRD data of P25 TiO₂ nanoparticles to the equation above, the particle size was evaluated at 18.1 nm. Furthermore, the ratio of anatase phase in those nanoparticles can be evaluated according to the following equation [36]:

$$A(\%) = \frac{100}{(1+1.265I_R/I_A)},\tag{9}$$

where I_R is the intensity of the peak at $2\theta = 27.5^{\circ}$ and I_A is the intensity of the peak at $2\theta = 25.3^{\circ}$ (those are the most prominent signals of rutile and anatase form, resp.). By using this equation, the anatase phase ratio in P25 TiO₂ is estimated at 84%.

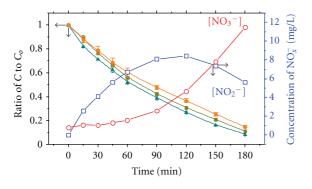


FIGURE 2: UV photocatalytic degradation of nitrobenzene at pH 4 (blue line with square symbols), 7 (green line with triangle symbols), and 11 (orange line with round symbols) with 0.2 g TiO₂ at 30°C and the air feeding of 150 mL min⁻¹ for 3 h. Concentrations of reaction intermediates nitrate (red line with empty circles) and nitrite ions (blue line with empty squares) at pH 7 are corresponded to the right ordinate as time changes.

3.3. Effects of pH on Photocatalytic Degradation of Nitrobenzene. The effects of pH on the photocatalytic degradation of nitrobenzene are shown in Figure 2. According to the degradation processes carried out in pH 4, 7, and 11, the most efficient degradation is in the case of pH 7 and the least one is in the case of pH 11. Because the surface of TiO₂ catalysts in aqueous solution is in the form of Ti-OH [37, 38], and the pH of zero point charge (pH_{zpc}) is about 6–6.4 for anatase phase and 5.8 for rutile phase [39]. In circumstance of pH value less than pH_{zpc}, the major species on the surface of catalysts in the form of Ti-OH₂⁺ is likely to occur whereas with a pH greater than pH_{zpc} the negative form of Ti-O⁻ is prone to happen. To produce strong performance of the photocatalytic reaction, it is important to have fair interaction between the substance to be degraded and the catalysts [40]. Therefore, the pH value of the solution plays a significant role in the interactions between the surface of TiO₂ and the instinctive electron property of the degrading substance. In the case of NB molecule, it is very likely to be in the form of positive charge because of the low pK_a for NB [41].

Another aspect of the pH effects that comes into play is the formation of 'OH radicals being more feasible in low pH condition [30], which is shown in Figure 3 that hydroxyl radicals generated at pH 4 has more amount than those in pH 7 and 11. According to the results shown in Figure 2, combining the effects cited above indicate the interaction between the substance and surface has more influence on the catalytical effect than the generation of •OH radicals. The additional data included in Figure 2 are the concentrations of intermediates NO₃⁻, NO₂⁻ which were generated during the degradation processes. The nitrite ions detected were assumed to be generated through the cleavage of nitro group on NB molecules. The nitrate ions were further produced through the oxidation of nitrite ions by the oxidants which were initiated by the UV irradiation of the catalysts and generated by a variety of reactions. These can be demonstrated from the observed trend in Figure 2 how the concentration of nitrites reached the peak at 120 min and

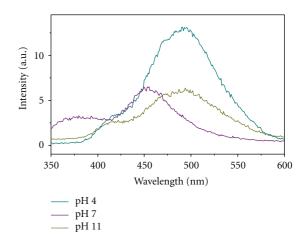


FIGURE 3: PL spectra of hydroxylated coumarin (7-hydroxycoumarin) solutions adjusted at pH 4, 7, and 11. Each 1.0 mM coumarin aqueous solution was mixed with 0.1 g TiO₂ and irradiated thereafter using an UV lamp with the peak intensity of 366 nm for 15 min.

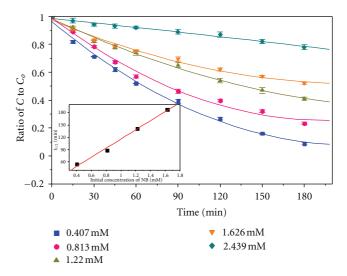
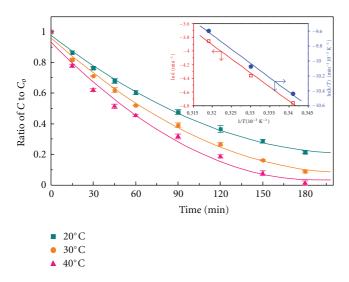


FIGURE 4: The initial concentration dependence of reaction rates for the photodegradation of nitrobenzene by TiO_2 as time changes. The processes were operated with 0.2 g TiO_2 and various concentrations of nitrobenzene at pH 7 and 30°C with the air feeding of 150 mL min⁻¹ for 3 h. Inset shows the effect of the initial concentrations on the half-life of NB.

since then nitrates increased at the expense of nitrites. This observation will be discussed along with the investigation of mechanisms in later section.

3.4. Physical Properties of the Photodegradation Reaction. Figure 4 exhibits the initial concentration dependence of reaction rates for NB photodegradation. To examine whether this reaction follows the Langmuir-Hinshelwood dynamic model, a graph of $t_{1/2}$ versus C_o was plotted according to the following equation [42]:

$$t_{1/2} = \left(\frac{0.693}{k_{\rm LH}K_{\rm ads}}\right) + \left(\frac{0.5}{k_{\rm LH}}\right)C_o,$$
 (10)



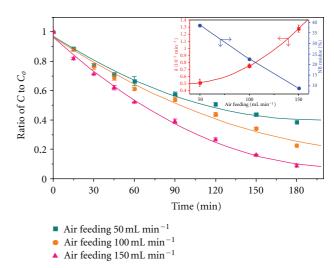


FIGURE 5: The temperature dependence of nitrobenzene photodegraded by TiO₂ as time changes. The processes were operated with 0.2 g TiO₂ and 0.407 mM nitrobenzene at pH 7 and 20°C, 30°C, 40°C with the air feeding of 150 mL min⁻¹ for 3 h. Inset shows the linear fits of the data for lnk and ln(k/T) versus the reciprocal of temperature.

where $t_{1/2}$ is the half-life of the reactant, k_{LH} is the rate constant of photocatalytic reaction, K_{ads} is the adsorption constant of reactant, and C_o is the initial concentration of the reactant. While the reactions were conducted under the condition of 0.2 g P25 TiO₂, at pH 7 and 30°C with the air feeding of 150 mL min⁻¹, by deriving from the slope and intercept of the linear fitted line, the following constants were acquired: k_{LH} is 4.48×10^{-3} mM min⁻¹, and K_{ads} is 36.02 mM^{-1} for NB, as shown in the inset of Figure 4.

The temperature dependence of reaction rates is shown in Figure 5, the reactions were carried out under the conditions of 0.2 g TiO₂ at pH 7 with the air feeding of 150 mL min⁻¹ for 3 h at temperatures of 20°C, 30°C, 40°C, respectively. According to the Arrhenius equation and the transition state theory [43], one can obtain the values of activation energy, entropy, and enthalpy of the reaction through the applications of these equations. From the left ordinate of the inset in Figure 5, the linear fit of the data is $\ln k = -4143.8(1/T) + 9.3549$ with the activation energy of 34.45 kJ moL⁻¹. The enthalpy and entropy of reaction can be evaluated from the linear fit of the data with the right ordinate $\ln(k/T) = -38.436 + 2.6511$: with $\Delta H =$ 0.31 kJ moL⁻¹ and $\Delta S = -0.041$ J moL⁻¹ K⁻¹.

3.5. The Air Feeding Dependence of the Photodegradation Reaction. The air feeding dependence of NB photodegradation by TiO₂ is shown in Figure 6. The processes were conducted with 0.2 g TiO₂ and 0.407 mM NB at pH 7 and 30° C for 3 h. The dissolved oxygen brought into the solution by air feeding could be used as the scavenger of the electrons on TiO₂ which were excited by UV irradiation. This can prevent the recombination of the photogenerated e⁻/h⁺ pairs since with such recombination, the efficiency of catalysts decreases dramatically. Furthermore, with large quantity

FIGURE 6: The air feeding dependence of nitrobenzene photodegraded by TiO_2 as time changes. The processes were operated with 0.2 g TiO_2 and 0.407 mM nitrobenzene at pH 7 and 30°C for 3 h. Inset shows the relationships between the air feeding and rate constant, NB residue, respectively.

of dissolved oxygen in solution, more oxygen radicals will be produced and as they are adsorbed on the surface of TiO₂, these radicals will react further with H₂O molecules to generate \cdot OH, which enhances the rate constants and therefore the efficiency of catalysts as shown in the inset of Figure 6. The results from the photodegradation in Figure 6 reveal that the larger amount of air feeding supplies the stronger performance it delivers, which is correspondent with the discussion above.

3.6. Investigation of the Mechanisms for NB Photodegradation Reaction. It has been proposed that the photodegradation pathway of NB is related to partial oxidation of NB to hydroxyl derivatives [44], such as *m*-, *o*-, *p*-nitrophenol. To investigate the mechanisms of this photodegradation reaction, the solutions were prepared as indicated in experimental section and were sampled at every half hour interval and up to 10 h of the reacting time span for HPLC separation times of HPLC authentic standards for the analytes, which is 6.4 min for *m*-nitrophenol, 8.7 min for *o*-nitrophenol, 5.7 min for *p*-nitrophenol, and 9.2 min for NB, the reaction intermediates can be identified with these data accordingly.

The concentration changes of the intermediates of NB photodegradation with corresponding time are shown in Figure 7, the total concentration of the intermediates is less than 7μ M, and it is conspicuous that *m*-nitrophenol was produced at the highest rate and *p*-nitrophenol the second highest for the initial 2 h of reaction and since then both decreased briskly while *o*-nitrophenol remained an approximate concentration level during this period. GC/MS analyses were also used to confirm the reaction intermediates. The pH decrease from the start of reaction of pH 5.84 to pH 3.82 after 3 h of reaction was observed as

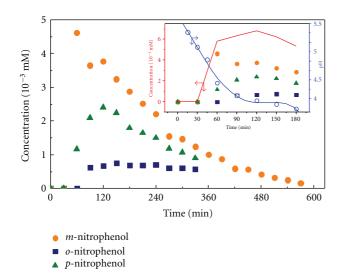


FIGURE 7: The concentration changes of intermediates m-, o-, p-nitrophenol as time changes. The photodegradation processes were operated with 1.626 mM nitrobenzene and 1.0 g TiO₂ at pH 7 and 30°C with the air feeding of 150 mL min⁻¹ for 10 h. Inset shows the relationship between pH changes and the total concentration of intermediates (in red line) for the initial 3 h of reaction.

shown in the inset of Figure 7. The pH value of *m*-, *o*-, *p*-nitrophenol aqueous solution of the same concentration was measured as 6.16, 5.74, and 5.98, respectively. Combining these results with the generation of nitrites and nitrates from Figure 2, it can be assumed that the NB photodegradation reaction involves hydroxyl group grafting on NB molecules and undergoing the nitro group cleavage from the molecules.

After the photodegradation reaction, 8.8% NB residue was observed from a system of 0.2 g TiO₂ and 0.407 mM NB at pH 7 and 30°C with the air feeding of 150 mL min⁻¹ for 4 h of photocatalysis whereas 4% residue from COD analysis was achieved with the same condition, it suggests that the mineralization of NB to CO₂ and H₂O may occur in parallel with the pathway cited above.

4. Conclusions

A new method of photocatalytic degradation of NB was proposed by coating the titania nanoparticles on a quartz tube through the introduction of tetraethyl orthosilicate into the matrix. Some parameters were discussed herein, such as the dependence of pH, temperature, air feeding on the photocatalysis, and some physical properties such as the activation energy, entropy, enthalpy, adsorption constant, and rate constant were also acquired. The photodegradation pathway was further investigated through HPLC, GC/MS, IC, and COD analyses.

The optimum efficiency of the photodegradation with NB residue as low as 8.8% was achieved with the experimental conditions mentioned. Although it is not the most efficient result compared to those from other studies, one of the advantages of practicing the NB photodegradation by adopting this approach is that the necessity of separating the catalyst from solution after use is not required and therefore making it easy to be applied repetitively. With these considerations, treating the wastewater from the industries by UV photocatalytic mineralization of organic pollutants using TiO_2 photocatalyst coated on quartz tube with tetraethyl orthosilicate as the binder is one of the convenient and potential candidates for this issue.

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